**Biomathematics and Statistics Scotland** 

# To carry out tern modelling under the Framework Agreement C10-0206-0387

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## **Joint Nature Conservation Committee**

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#### CONTENTS

1.	Non Technical Summary	3
2.	Introduction	4
3.	Data	5
4.	Methodology	8
5.	Results	14
6.	Discussion	19
7.	Maps Appendix	20

In addition to this report, there are two further documents associated with this project:

(i) BioSS Terns Report – Results Appendix;

(ii) BioSS Terns Report – Software;

and also ancillary files:

(iii) Spreadsheet files of grid predictions for each of the twelve species/colony combinations;

(iv) R code files for: data exploration; model fitting; grid predictions

(v) cleaned and standardised versions of the original data files.

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#### **1. Non-Technical Summary**

The Joint Nature Conservation Committee (JNCC) is working on the identification of important marine areas around the UK that are used by five species of tern during the breeding season. For the four larger tern species (Arctic, common, roseate and Sandwich terns), data are available from boat surveys, using both visual tracking and transect survey methods.

Following a competitive tendering process, in December 2011 BioSS was tasked with analysis of the visual tracking data for the four larger species of tern.

The analysis was to use the observed tern data in combination with data simulated from the whole of the sample area, with statistical models attempting to relate the locations of terns with background environmental variables. From these statistical models, predictions were to be made on a grid of data and then mapped, showing areas which are preferred by terns.

We found during analysis that the existing statistical methodology was inadequate for analysis, and that a procedure which accounts for the fact that birds are tracked for different lengths of time was needed.

The environmental data contained sets of variables which were very highly related to each other, and so the analyses can only assess associations, rather than identify drivers. We found that for most species, the most important predictor was the distance to the colony itself, but above and beyond that, other variables were also important. For both the Arctic and common terns, the average sea surface temperature in spring proved an important predictor. With only one colony of roseate terns, we could not draw conclusions of consistency across colonies, but chlorophyll may play a part in addition to distance to the colony and sea surface temperature. For Sandwich terns, the clearest predictor was distance to shore, and while other variables seemed important for particular species, there seemed to be nothing consistently so.

#### 2. Introduction

The problem at hand is thus: given the locations of terns, recorded by tracking (as described in Section 3), how can we learn about the preference of the birds in respect to how they select where to forage (based on the explanatory environmental variables we have), and how can this information be used to make predictions about locations frequented by different tern species?

As we are not sure as to the kind (shape) of relationships between tern preference and the environmental variables – that is, whether they will be linear or non-linear – we will need to consider a class of models which allow flexible regression relationships (Generalised Additive Models, GAMs) as well as those which are less flexible (Generalised Linear Models, GLMs).

The first difficulty is in the form of the data. Since birds fly around and since we wish to distinguish between foraging and commuting behaviour, we cannot simply take a random sample of locations and assess whether these locations were visited by terns, which would be a straightforward logistic regression analysis. Instead, the data represent, in essence, "presence-only" data. Ecologists, taking a lead from Boyce and McDonald (1999), who in turn borrowed from the medical statistics literature, started using so-called case-control designs; in medicine, these studies compare individuals having a condition (the "cases") with "control" individuals who do not have the condition. In spatial ecology, this equates to the observed presences being the cases, but unless absence can be determined reliably (as might be possible, for example, if we were studying nesting sites), we lack controls. A partial solution has been to sample from the entire possible range for the population at hand, and to use these as controls; this represents a so-called use-availability design. These controls represent pseudo-absences rather than true absences and because the number of controls is determined arbitrarily, we can only estimate relative preference and not absolute preference.

The second problem is that the data consist of repeated observations for a single bird, and that the number of repeats can vary enormously, depending on the length of the trip each bird made and how long the observers were able to track it. If not accounted for properly, this can bias the results. However, thinning the data set to a single observation per individual is not sensible, given the amount of data loss involved and the fact that we are interested in location-specific covariates, although a certain amount of thinning (equally applied to all individuals) was necessary to make the data set manageable for analysis. Previous work in the area (e.g. Aarts *et al.*, 2008) applied "mixed models" to account for between-individual variation, but does not seem to have fully addressed the problem of serial correlation within a track (see later). The tern data set differs from the one studied by Aarts *et al.* (2008) in that we have only one track per individual, so we cannot distinguish variability between individuals from variability between trips. Thus rather than simply applying existing methodology, there was also to be an element of method development.

Another difficulty relates to the spatial nature of the data. Technically there is a temporal element too, but this only applies within a single track, so we consider the spatial aspect. It is well understood (Beale *et al.*, 2010) that failing to deal with spatial autocorrelation – that is, the fact that data from locations close in space are typically more similar than data further apart – can lead to incorrect statistical inferences; in particular, significant associations can be claimed which are in fact spurious. Modelling spatial autocorrelation can be tricky, especially for the class of binary data we will be dealing with. This ultimately required the use of software which is currently undergoing development in the R statistical package, the software of choice for this project.

The R software (R Development Core Team, 2012) contains numerous packages for fitting the binary response data we have. Our task is to experiment with different models and packages to obtain appropriate functions for fitting the kinds of models required to suit our data. As our background environmental variables are themselves highly cross-correlated, finding important associations can be difficult. Appropriate application of model selection methods is one strategy for reducing multicollinearity, although multicollinearity is less of a problem when the purpose is prediction rather than explanation (Shmueli, 2010). For this reason we will want to compare a range of model selection options.

#### 3. Data

The data received from JNCC consisted of locations where tracked birds were recorded to be foraging; locations where birds were commuting were removed from the data prior to analysis. Foraging observations were thinned to every 10<sup>th</sup> observation prior to analysis. Environmental covariates provided are shown in Table 1; all except sediment, which is a categorical variable used to derive sand, were considered as potential covariates in the analyses.

Parameter	Data set	Source	Date collected	Processing	Original scale and projection	Data type
Chl_month	Chlorophyll-a concentrations, mg/m <sup>-3</sup> , monthly	PML	2009	Imagestakenat1.2kmsquare,re-mappedto1km square	Approx 1.1km Transverse Mercator	Continuous average concentration values by month
Dist_col	Distance to nearest con-specific colony (m)	Nearest colony identified from JNCC tern colony maps	N/A	Hawths Tools distance between points function in ArcGIS	1km <sup>2</sup> grid cells OSGB 1936 Transverse Mercator	Continuous distance values (metres)
Dist_ shore	Distance to nearest mainland coast (ie shortest distance to coast)	Nearest coastline identified from an Ordinance Survey high water polygon	N/A	Joins and Relates in ArcMap to store distance to closest shore for each point in the environmental layers grid.	1km <sup>2</sup> grid cells OSGB 1936 Transverse Mercator	Continuous distance values (metres)
Sal_spring	Sea surface salinity in spring (‰)	Proudman Oceanographic Laboratory	10 year simulation	Bilinear interpolation, and Inverse distance weighted interpolation to fill in missing values near the coast	0.01 <sup>2</sup> decimal degrees GCS WGS 1984	Continuous salinity values derived from simulation of POLCOMS
Sal_summ	Sea surface salinity in summer (‰)	Proudman Oceanographic Laboratory	10 year simulation	Bilinear interpolation, and Inverse distance weighted interpolation to fill in missing values near the coast	0.01 <sup>2</sup> decimal degrees GCS WGS 1984	Continuous salinity values derived from simulation of POLCOMS
sst_month	Mean surface temperature by month (°C)	PML	2006- 2010	Imagestakenat1.2kmsquare,re-mappedto1km square	Approx 1.1km Mercator	Continuous average temperature values by month

#### **Table 1 – Environmental covariates**

Strat_temp	Surface to seabed temperature difference in summer (°C)	Proudman Oceanographic Laboratory	10 year simulation	Bilinear interpolation, and Inverse distance weighted interpolation to fill in missing values near the coast	0.01 <sup>2</sup> decimal degrees GCS WGS 1984	Continuous temperature difference values derived from simulation of POLCOMS
Northness_1s	Seabed aspect from -1 (south) to 1 (north)	Derived from Defra digital elevation model data.	NA	Aspect function followed by transformation to radians and trigonometric cosine function, in ArcGIS Spatial Analyst	Approx. 30m <sup>2</sup> grid cells, varies slightly across extent of data. GCS WGS 1984	Continuous values from -1 to 1.
Eastness_1s	Seabed aspect from -1 (west) to 1 (east)	Derived from Defra digital elevation model data.	NA	Aspect function followed by transformation to radians and trigonometric sine function, in ArcGIS Spatial Analyst	Approx. 30m <sup>2</sup> grid cells, varies slightly across extent of data. GCS WGS 1984	Continuous values from -1 to 1.
Bathy_1s	Seabed depth (m below lowest astronomical tide)	Defra digital elevation model.	NA	Triangulation with linear interpolation	Approx. 30m <sup>2</sup> grid cells, varies slightly across extent of data. GCS WGS 1984	Continuous depth values
Sediment_250	Seabed sediment/substrata	British Geological Survey (DigSBS250)	NA	Simplification of DigSBS250 Folk categories supplemented by additional data	Vector dataset GCS WGS 1984	Dominant sediment type categories: mud and sandy mud, sand and muddy sand, mixed sediments, coarse sediments, rock
Sand	Seabed sediment: 'sandy' category (1) or 'other' category (0)	Based on Sediment_250 with folk triangle.				

Slope_1s_deg	Seabed slope (° incline between adjacent grid cells)	Derived from Defra digital elevation model data.	NA	Slope function in ArcGIS Spatial Analyst	Approx. 30m <sup>2</sup> grid cells, varies slightly across extent of data. GCS WGS 1984	Continuous slope values
Ss_ currents	Shearstress:currentsMaximumtidalforce(Newtons/m²)	Defra funded Plymouth Marine Laboratory project	June- August from 1998 to 2008.	Inverse distance weighted interpolation, derived from proWAM 12km wave model	0.003 <sup>2</sup> decimal degrees GCS WGS 1984	Continuous values
Ss_ waves	Shearstress:wavesMaximumwaveforce(Newtons/m²)	Defra funded Plymouth Marine Laboratory project	June- August from 1998 to 2008.	Inverse distance weighted interpolation, derived from POLCOMS model.	0.003 <sup>2</sup> decimal degrees GCS WGS 1984	Continuous values
Spring_front	Probability of a frequent thermal front in spring. Ratio of strong thermal fronts to observations, averaged over all years.	Defra funded Plymouth Marine Laboratory project	June- August from 1998 to 2008.	Bilinear interpolation	Approx 1.2km <sup>2</sup> GCS WGS 1984	Probability from 0 to 1.
Spring_frt_sd	Interannual standard deviation of probability of a frequent thermal front	Defra funded Plymouth Marine Laboratory project	June- August from 1998 to 2008.	Bilinear interpolation	Approx 1.2km <sup>2</sup> GCS WGS 1984	
Summ_front	Probability of a frequent thermal front in summer. Ratio of strong thermal fronts to observations, averaged over all years.	Defra funded Plymouth Marine Laboratory project	June- August from 1998 to 2008.	Bilinear interpolation	Approx 1.2km <sup>2</sup> GCS WGS 1984	Probability from 0 to 1.
Summ_front_sd	Interannual standard deviation of probability of a frequent thermal front	Defra funded Plymouth Marine Laboratory project	June- August from 1998 to 2008.	Bilinear interpolation	Approx 1.2km <sup>2</sup> GCS WGS 1984	

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#### 4. Methodology

This section provides brief details of and justification for the methodology used during this project.

#### 4.1 Case-Control Design

We use a case-control approach as described in Aarts et al. (2008). It should be noted that this is actually a use-availability design (Keating and Cherry, 2004) rather than a case-control design, since the controls represent pseudo-absences rather than true absences. Logistic regression is used to model a response variable which takes the value 1 for the observations and 0 for the control (available environment) points. The exponential function of the linear predictor is then proportional to the expected density of observations (Aarts, 2012). Warton and Shepherd (2010) demonstrate that in the case of pseudo-absences that are regularly spaced or located uniformly at random over the region, the logistic regression slope parameters (but not the intercept) converge to those of the corresponding inhomogeneous Poisson point process model as the number of pseudo-absences increases. The observed data set does not consist of all locations where terns forage, or even all locations where terns are foraging at one specific point in time; instead it is locations where terns were recorded to be foraging. The logistic regression approach models the probability that a point is a presence not a pseudo-absence. This probability has no physical meaning and tends to zero as the number of control points increases; it is the intensity of the presences rather than the probability of occupancy that is of interest and a Poisson point process model therefore has a more natural interpretation. Existing software does not allow a spatial Poisson point process modelling approach to the current problem to be fitted routinely. We therefore take the approach of generating control samples and using logistic regression to approximate the point process model. However, future developments in the INLA (integrated nested Laplace approximation) package (INLA, 2012) should allow the point process model to be fitted directly without the need to generate control samples.

As we wanted to consider a range of modelling options, we proposed generating the control samples so that each set resembled the foraging locations in the original case data (particularly in relation to exhibiting autocorrelation). The simplest way of implementing this seemed to be to generate sets of initial "track starts" – one for each track in the case data – and then applying the rest of the case track movements between foraging locations to each. If a control "track start" fell on land, it was replaced with another random starting location; subsequent points falling on land were simply omitted. This, in effect, creates a set of control samples which are the original case tracks relocated randomly throughout the suggested range of the species – see Figure 1. Having control tracks (rather than just control points as per Aarts et al., 2008) allowed us to think sensibly about using different forms of random effect structure when modelling. However, given the approach eventually chosen did not include random effects, future work could revert back to using control points rather than control tracks; these will provide better spatial coverage per sample point than tracks (therefore reducing the sample size and processing time needed), but will not enable small-scale assessment of spatial coverlation.

Control locations (the initial track starts) were generated by taking a random angle (uniformly distributed between 0 and  $2\pi$  radians) and then a random distance from the colony (also uniformly distributed between zero and the maximum foraging range). This gives a greater density of control points closer to the colony than if points were uniformly distributed over a circle. Aarts *et al.* (2008) recommend that control points should be selected in proportion to accessibility; this would certainly mean a greater density closer to the colony, although it does not necessarily imply a uniform distribution of distance. Selecting points in proportion to accessibility means that the model outputs

provide estimates of preference, which is defined as the ratio of use to availability. Estimates of usage can then be obtained by multiplying preference by accessibility. On the other hand, if the control points were uniformly distributed over the circle this would mean that the model outputs would provide direct estimates of usage. The number of control points needs to be sufficiently large to ensure that estimates of slope parameters converge to stable estimates. Exploratory analysis by JNCC suggested that up to five or six control tracks may be required. To be cautious, twelve replicate control tracks were generated for each case track. Further research is needed to determine the number of control points required to achieve stability for different sample sizes as this will not necessarily be a multiple of the number of observations. This is likely to be a complex problem, as the number of control points required could potentially vary hugely between colonies – it will be a function of colony size, colony density, topography of nearby land masses, interactions with other species, availability and distribution of food sources, and so on.



Locations: Coquet\_Roseate\_10.csv

Figure 1 – Example (for roseate tern at Coquet) of the pseudo-absences (in black) generated by shifting the original track (in red)

The environmental information described in Table 1 was attached to each case and control point. To make grids for interpolating the preferences (and usage) to the whole foraging range of each species at each colony, environmental information for points spaced every 500m were provided for the area within the maximum foraging range for each species, around each colony. This was either 30km (Arctic, common and roseate terns) or 60km (Sandwich terns).

#### 4.2 Statistical Model and R Packages

The modelling needs to take account of the fact that points on a given track are not independent (i.e. they are repeated measures), as birds generally move only a short distance from one observation to the next. Failure to account for the lack of independence between observations within track – a form of pseudoreplication – leads to underestimation of the variance of parameter estimates and might therefore result in some environmental variables being wrongly regarded as significant. The simplest way to deal with this is to weight each observation by the reciprocal of the length of the observed track. This has the effect of treating each track as a single sampling unit, instead of treating the individual observation as the sampling unit, but is preferable to averaging data over tracks as individual covariate values are retained for each point. Thus, the basic statistical model used in our analysis is a weighted binomial generalised linear model (GLM) or generalised additive model (GAM) with a logit link. As the data are binary the dispersion parameter should be fixed at 1.

Existing methodological approaches (for example, Wakefield et al., 2011) fit a generalised linear mixed model (GLMM) to account for between-individual variation, although it is unclear exactly how these authors deal with the control points when specifying the random effects. We considered several approaches to the tern analysis using random effects. The standard approach to repeated measures would be to fit a separate random effect for each case or control track. However, this is not appropriate in our situation, because the case tracks consist entirely of presences while the control tracks consist entirely of absences, leading to biased estimates of the fixed effects of the covariates. An alternative approach that we considered, which was probably the one used by Wakefield *et al.*, is to fit a random effect for each combination of a case track with its corresponding control tracks. Unfortunately, this approach is not suitable either, as within each level of the random effect there are units from the same control set that are highly correlated and units from different control sets which are uncorrelated; the model incorrectly assumes that these correlations are equal. For smaller numbers of control sets many of the entries in the correlation matrix are between units from the same track, and as in the previous approach parameter estimates are biased. It is likely that a variation of such a model accounting for spatial autocorrelation would help reduce this bias. As the number of control tracks increases, so entries in the correlation matrix that correspond to units in different control sets start to dominate; this means that the variance of the random effect tends towards zero as the number of control sets increases and the results approach those for an unweighted GLM that does not account for dependence within a track/control set combination. As described above this gives unbiased estimates of the parameters, but estimates of the standard errors are biased (i.e. the parameter estimates falsely precise) because they fail to take account of the dependence within a track. Under-estimation of the standard errors would have been even greater if we had used control points instead of control tracks.

We felt that none of the approaches we explored for specifying the random effects were appropriate, thus we chose to use a fixed effects only weighted model. Weighting avoids biasing the results towards the longer tracks, but would still be necessary even if all the tracks were of the same length because of the autocorrelation between observations in the same track. As there is not time for birds to fly far between successive foraging observations and as the environmental covariates are spatially autocorrelated, observations from the same track tend to occur in similar environmental conditions. This does not necessarily indicate differences in individual preference between birds; multiple tracks per bird would be needed to investigate this.

Our model exploration work led us to three different functions within R, representing different levels of complexity, but all fitting variants of the same basic model:

- 1. Function glm() in the base "stats" package;
- 2. Function gam() in the "mgcv" package (Wood, 2011); and
- 3. Function inla() in the developmental INLA package (INLA, 2012).

All three functions can fit the weighted binary logistic regression required. The first function, glm(), is the GLM workhorse of R, is extremely stable and reliable and has a large suite of generic helper functions. The function gam() in package mgcv provides a simple, flexible interface for fitting smooth functions for the effect covariates, in case some relationships are more complex than straight lines. Finally, we looked for a suitable method to account for residual spatial autocorrelation (i.e. the remaining autocorrelation after the effect of the environmental covariates has been accounted for). Not accounting for this autocorrelation can lead to underestimation of the standard errors for parameter estimates, and this in turn implies some variables may wrongly be declared significant as a consequence; for more detail see Beale et al. (2010). The track structure of the data means that a simple rectangular grid would not be appropriate, and the data sets are too large, in general, for modelling large covariance matrices based upon distances between observations (note standard methods consider all pairwise differences). The inla() function in the INLA package solves this problem by defining a "mesh" to represent the spatial autocorrelation structure based upon the locations of the observations; in this way, one observation is linked to a small number of its closest neighbours, and the modelling is intelligent enough to account for these between-neighbour distances. Happily, this method has allowed us to fit full spatial models with whole-species data sets. For an example of the mesh structure produced by the INLA package, see Figure 2.

In summary, the analysis takes three distinct stages; firstly, we fit generalised linear models and perform model selection. The high level of correlation between explanatory variables suggests that we need to be careful about model selection and use different methods depending on whether the aim is prediction or explanation. Because of this concern, we in fact use three different approaches to selection on the basis of: AIC (Akaike's Information Criterion); BIC (Bayesian Information Criterion); and significance of individual terms via likelihood ratio tests (LRTs). The AIC is known to produce large models, with many explanatory variables which may not be statistically significant, but is optimal in terms of linear models for prediction. The BIC contains a penalty (against adding variables to a model) which is stronger than that for AIC and which is a function of number of observations. BIC is consistent in the sense that if the true model is among the candidate models then the probability of selecting it approaches 1 as the sample size increases, but it is sub-optimal for prediction. As our "number of observations" (which includes the pseudo-absence data) is typically very large, the penalty is itself very large and hence the models selected by BIC tend to be much smaller than those chosen by AIC.

#### **Constrained Delaunay triangulation**



Figure 2 – Mesh to represent spatial correlation structure of response data; example shown for Forvie colony and Sandwich terns

The LRT method is preferred for identification of important associations between tern presence and the covariates, as each term in a selected model must be individually statistically significant; we use a significance level of 10% here as we wish to explore feasible models thoroughly. Note, however, that if two covariates are highly correlated and both are included in the model, neither may appear significant, even though either one of them would be significant if fitted individually. Even in the absence of correlation between the potential covariates association does not necessarily indicate causality as the significant variables may be correlated with a missing, causal, covariate. At this point we also check for consistency of covariate effects across different years, for those species/colony combinations which have more than one year's worth of data; in essence, we fit interaction terms between year and the LRT-selected covariates and conclude effects are not temporally consistent if the interaction is significant.

These generalised linear models will not account for non-linear relationships between response and covariates. For this, we fit generalised additive models (GAMs) which use "spline functions" to

describe these response curves – these are non-linear, non-parametric functions which can take any shape, limited only by a number of "basis functions" which control how "wiggly" the line is allowed to be. We chose three basis functions so that the response-covariate relationships would be fairly smooth – we do not believe *a priori* that the relationships should be too complex. For GAM models, as the AIC is only approximate, model selection was instead performed by assessing the significance of smooth terms by comparison to a chi-squared distribution. Next, the INLA method is used to assess the effect of any spatial autocorrelation in the response data, over and above that explained by the environmental covariates in the model. INLA is applied to the model chosen by the GAM model selection; most models chosen across the species/colonies contained some smooth terms, and we wished to have a consistent approach. INLA allows us to account for spatial autocorrelation within the model and to identify those parameters which may have been falsely identified as significant.

Finally, we use the fitted models to produce maps of the estimated preferences and estimated usage, as detailed in Aarts *et al.* (2008). Again, for reasons of consistency we aimed to use the selected GAM model where possible, although variables which the INLA model suggests are no longer significant after allowing for spatial autocorrelation could be removed. The maps of preference show which "habitats" the terns might prefer, to a degree, but note this interpretation is not strictly applicable if either "distance to colony" or "distance to shore" are in the selected model.

If there was no preference for particular habitats, the odds ratio would be equal to the ratio of the number of observations to the number of controls. To calculate preference the odds ratio is adjusted by multiplying by the number of controls per observation, which is equivalent to taking the exponential transformation of the linear predictor and multiplying by the number of controls per observation. This does not purely reflect preferences for environmental variables because of bias in our accessibility model, which assumed a uniform distribution of distance from colony. To correct for this the linear predictor could be broken up into two parts: a component due to distance from colony and a component due to environmental variables alone.

The maps of usage adjust for accessibility, using the same model as that for generating the control sample locations - i.e., a uniform distance distribution from the colony. Equation (5) of Aarts *et al.* is:

$$f_u(X_s) = \frac{h(X_s)}{1 - h(X_s)} r f_a(X_s)$$

where  $f_u(X_s)$  is the spatial probability density function for usage,  $h(X_s)$  is the predicted value from the fitted model, *r* is the number of controls per observation and  $f_a(X_s)$  is the probability density function for accessibility. In our case

$$f_a(X_s) \propto \frac{1}{\text{distance}}$$

so preference is divided by distance to colony and multiplied by a scale factor which ensures that the probabilities sum to one. These results could then be multiplied by the number of birds in the colony.

#### 4.3 Verification of the Weighted Model

How do we know the weighted regression model is appropriate? This is easily verified by thinking about what the model would look like if there were only a single observation per bird, rather than the tracks that we have currently. There would then be no basis for random effects or weights, and we could fit a simple logistic regression (preferably accounting for spatial autocorrelation in some way – but that is not relevant to the point we're trying to make so we'll ignore it for now). We have tested the "single point" model by sampling randomly a single point from each track, and then running this simple analysis. By generating multiple single sets, we were able to gauge how variable the "single point" results were. Although the results will be quite variable, they provide the perfect yardstick for the "full track" analysis. The weighted logistic model parameter estimates were very similar to the mean estimates from the single point analysis and the standard errors were also similar, as was to be expected, since the weights are effectively accounting for the track lengths; in addition the results were very insensitive to the number of control sets used.

Note that the "single point" analysis represents the maximum amount of thinning that could be applied to the data. In consequence, the single-point analysis described above also (at least partially) resolves the issue of thinning; as long as the data are not thinned too much – where "too much" would imply that possible relationships with covariates are lost, perhaps by obtaining a track on a coarser scale than the covariates – the actual level of thinning is unimportant.

#### 4.4 Other Modelling Approaches / Packages Explored

When the use of a mixed effects modelling approach was investigated initially (before it was found unsuitable), a variety of functions /packages were explored. For fitting a GLMM, both glmer() in package lme4 and inla() again appeared good possibilities. The function glmmadmb in package glmmADMB did not perform well and, in any case, does not appear to allow a weights option. Function glmmPQL in package MASS was not suitable for two reasons: the PQL (penalised quasi-likelihood) fitting process is known to produce biased results; and there does not seem to be a possibility of fixing the dispersion parameter. The gamm() function in mgcv appears to use PQL and therefore suffers from the same problems as glmmPQL.

#### 4.5 Discussion

Our tender suggested exploring a mixed models approach, as this appeared to have been successful in earlier work (e.g. Aarts *et al.*, 2008; Wakefield *et al.*, 2011). In Aarts *et al.* (2008) there are multiple observed tracking instances of the same bird – this would seem to be the difference between their data and the current terns data, and may explain why a mixed model approach may have worked for them and not for us. However, it also appears that these earlier studies may have ignored the fact that whereas observations within a track are serially correlated, there is no correlation between an actual observation and the corresponding control observations. This may have led to standard errors being under-estimated in these studies. We might speculate that a weighted analysis may have been beneficial in these earlier works as well as with our current analysis.

#### 5. Results

This results section contains a summary discussion of the results obtained for the analyses of the twelve data sets. This section is ordered by colony within tern species. The full R output for each model is contained in the "Results Appendix" accompanying this report. The sections in that document mirror those here – for example, section 5.1.1 below summarising the Coquet Arctic tern results is mirrored in section A.1.1 in the Results Appendix document. This should enable straightforward cross-referencing where required.

Note that the models presented are the ones chosen via model selection procedures. Owing to the very large correlations between the environmental covariates, it would be wrong to infer *causal* relationships from the *correlative* relationships discovered; we are finding association rather than proving what is driving habitat selection. Nonetheless, this does not preclude obtaining good predictions. In other words, association rather than causation can provide the information we require to assess which areas are suitable for foraging behaviour.

#### 5.1 Arctic Terns

### 5.1.1 Coquet

This analysis illustrates the harsher penalty from the BIC for model selection; that produces a model containing only distance to colony, whereas AIC selects distance to colony, sst\_may, summ\_front\_sd, ss\_wave and ss\_current. LRT selects the same model as AIC, and the effects across the three years' worth of data were consistent.

The GAM model selected also uncovered distance to colony, sst\_may, summ\_front\_sd, ss\_wave and ss\_current; an illustration of the relationships is shown in Figure A.1.1.4. From this it can be seen that preference increases with increasing values of sst\_may and ss\_current, and decreases with increasing values of ss\_wave and summ\_front\_sd, although the relationship with sst\_may seems to be largely driven by a small number of outliers. Of those four, the relationship with ss\_wave seems strongest (p=0.0125). The relationship with distance to colony shows that preference decreases steadily as you move further away from the colony (to about 20km) and then the relationship levels off. This is the strongest effect of all (p<0.001).

The INLA model suggests that spatial autocorrelation may be causing the suggestion of significance for some of the variables noted above; from the INLA output, we have credible intervals rather than p-values, and we regard a variable as important if the interval does not overlap zero. In fact, only sst\_may (0.0994, 3.2528), ss\_wave (-0.0804, -0.0129) and the smooth distance to colony relationship remain as important covariates once spatial autocorrelation is taken into account. However, the smooth relationship for distance to colony contains a seemingly unlikely upturn for large distances, likely an artefact caused by a lack of case data at such distances (see Figure A.1.1.5).

#### 5.1.2 Farnes

Here AIC selects distance to colony, sst\_april and summ\_front\_sd, while BIC leaves out sst\_april of the three. LRT again agrees with AIC. With only one year's worth of data, there was no need to check for consistency across years. GAM also selected distance to colony, sst\_april and summ\_front\_sd, and as seen in Figure A.1.2.4 the relationships with distance to colony and sst\_april

are decreasing whereas that with summ\_front\_sd suggests higher values correspond to higher preference. For this data set, the INLA analysis supports the model selected as none of the credible intervals overlap zero. The covariate relationships were sufficiently linear for INLA to dispense with the GAM element.

#### 5.1.3 Outer Ards

AIC selects distance to colony, distance to shore, chl\_apr, chl\_jun and sst\_may; BIC on the other hand finds only chl\_apr and sst\_may, ignoring the distance variables completely. The LRT selection adds distance to colony to the BIC selection. There was no evidence of covariate effects being different across years. Interestingly, the GAM model selects distance to colony, chl\_apr and sst\_april, differing from the LRT selection by switching sst\_may and sst\_april. Figure A.1.3.4 shows that for all three selected GAM covariates, higher values are associated with lower preference. The spatial INLA analysis provides three credible intervals, none of which overlap zero, so there is no evidence on this occasion that any of the covariate effects found are due solely to spatial correlation. Again, the covariate relationships were linear and so INLA ran a spatial GLM rather than a spatial GAM.

#### 5.2 Common Terns

#### 5.2.1 Coquet

Here AIC selects distance to colony, chl\_june, sst\_april and ss\_wave, but as with the Arctic terns, BIC selects only distance to colony. The LRT model selection matches the AIC choice, and note here there is very slight evidence of the effect of distance to colony varying across years; however, we note that p=0.07443 and hence the evidence is weak; we assume the effects are consistent across years, therefore. The GAM model chosen contains the same covariates as the AIC and LRT selections. Figure A.2.1.4 suggest two linear relationships, with chl\_june (increasing) and ss\_wave (decreasing), and two non-linear relationships. With distance to colony, the relationship is clearly decreasing, but the rate of decrease increases further away from the colony. The relationship with sst\_april is also decreasing, but perhaps levelling out for the upper end of the range of temperatures. The spatial INLA analysis does not contain any intervals overlapping zero, so this reaffirms the significance of the four selected variables. The plot for sst\_april in Figure A.2.1.5 is interesting; it suggests that the negative slope is due to the outlying values of temperature at the bottom end of the range, and that in fact the relationship is an increasing one for the bulk of the data in the region around 7 °C. Further work should consider re-running the model excluding these outliers.

#### 5.2.2 Larne Lough

AIC selects distance to colony, distance to shore, sst\_april and bathy\_1sec. BIC finds three of these, dropping sst\_april. LRT and GAM agree with AIC. There is no evidence of these covariate effects being different across years. The relationship with distance to colony is negative, whereas for distance to shore and sst\_april the relationships are positive (see Figure A.2.2.4). The relationship with bathy\_1sec is "cup-shape", having lowest preference in the -150 to -100 metres range, but higher when deeper than 150 metres or shallower than 100 metres. The spatial INLA analysis supports the importance of all the variables except bathy\_1sec; note the very wide interval estimates in Figure A.2.2.5.

#### 5.2.3 Leith

For Leith AIC selects a large number of variables: distance to colony, distance to shore, chl\_may, chl\_june, sst\_may, summ\_front, spring\_front, sal\_spring, bathy\_1sec, slope\_1s\_deg. However, the large size of the data set means that the penalty on adding variables with BIC is huge, and therefore only distance to colony is selected. The selection with LRT is large, but not quite as large as for AIC; the variables here are distance to colony, distance to shore, chl\_may, chl\_june, sst\_may, sal\_spring, bathy\_1sec and slope\_1s\_deg. With so many variables, it is not surprising that one (sst\_may) displays weak evidence (p=0.0604) of varying between years, but again we choose to assume a consistent effect across years. (Note that the consequence of not doing so means that we can then only make predictions on a per-year basis, and only for the years for which we have data.)

The GAM model selected matches the LRT model. It can be seen clearly in Figure A.2.3.4 that most of the variables selected have quite small effects on preference relative to the effect of distance to colony, by far the strongest effect (p<0.001). The variables chl\_may, chl\_june, sst\_may and sal\_spring all have very small positive relationships with preference. The effect of slope\_1s\_deg is a little greater, although there is uncertainty for high values owing to the relative sparseness of data. Distance to shore shows a moderately negative association with preference. There is a slight non-linear relationship for bathy\_1sec, with an increasing relationship for lower depths levelling out for higher values. Distance to colony has a negative relationship which tends to get steeper for higher distances.

The spatial INLA analysis suggests the relationships with chl\_may and sal\_spring may simply be due to spatial correlation in the response data, as their credible intervals overlap zero, (-0.0007,0.0314) and (-1.6119, 2.0057) respectively.

#### 5.2.4 Mull

For Mull, the AIC selected chl\_apr, chl\_may, sst\_april, sst\_may and ss\_wave. BIC however selected only chl\_apr. LRT was different again, finding chl\_apr, chl\_may and ss\_wave. And GAM was different too – the variables here were chl\_apr, chl\_may and bathy\_1sec. It should be noted here that the Mull colony appears to be surrounded by a lot of land on all sides, with the sea (or lochs) forming narrow strips around the colony. It is not too surprising in particular that distance to colony is not significant – there are areas of water close to the colony which have land in between; it may not have always been possible to continue tracking terns which flew over such land, hence a number of possible areas close to the colony "as the tern flies" may not have been sampled, and thus may have caused bias in the results. We do note that chl\_apr and chl\_may do occur consistently in chosen models.

Figure A.2.4.4 suggests that all three relationships from the GAM model are non-linear. For both chl\_apr and chl\_may, preference increases with increasing chlorophyll for low concentrations, then levels off for higher values. For bathy\_1sec the relationship is generally negative, levelling out or even increasing slightly for higher values (shallower depths). The spatial INLA analysis broadly supports the GAM model.

#### 5.3 Roseate Terns

#### **5.3.1 Coquet**

The AIC chooses a large model here; the variables are distance to colony, distance to shore, chl\_may, sst\_june, sst\_may, summ\_front\_sd, strat\_temp and ss\_current. The BIC model is much smaller, having only distance to colony, sst\_june and sst\_may. The LRT model is in between, being composed of distance to colony, chl\_may, chl\_apr, sst\_june, sst\_may and summ\_front\_sd, and there was no evidence of differential year effects. The GAM model selected matches the LRT model. As shown in Figure A.3.1.4, the effects are all linear, with the strongest being for sst\_june and sst\_may (both p<0.001), although the relationship is positive for sst\_may but negative for sst\_june. There are weaker positive relationships with chl\_apr and summ\_front\_sd, and weaker negative relationships with distance to colony and chl\_may.

For the spatial INLA analysis, the credible interval for chl\_apr is (-0.0034,0.6158), which (only just) overlaps zero, casting doubt on the importance of that variable in terms of association with preference.

#### 5.4 Sandwich Terns

#### 5.4.1 Coquet

For Coquet, the AIC chose distance to colony, distance to shore, chl\_june and sst\_april. BIC only chose the two distance variables. LRT also chose distance to colony, distance to shore, chl\_june and sst\_april, and spotted a potential cause for concern re: interactions with year – the interaction of distance to shore and year was marginally significant with p=0.04994. Again we choose to ignore this interaction, so the maps produced (and presented in Section 7) work on the basis of average effects across all years (produced from the model without interaction term).

The GAM model matches the AIC and LRT models. Figure A.4.1.4 shows the four relationships, which all have different characters. The relationship with chl\_june is very small and positive. That between preference and sst\_april is non-linear, being flat for lower temperatures but decreasing for higher temperatures. There is a reasonably strong negative relationship with distance to shore, but the strongest effect of all (if not the most significant) is for distance to colony, generally decreasing, but faster with longer distances. Note that it is the high uncertainty with the distance to colony effect – characterised by the wide interval estimates in Figure A.4.1.4 – that cause distance to colony to have a higher p-value than distance to shore, despite having ostensibly a stronger relationship with preference.

With the spatial INLA analysis, the credible interval for chl\_june overlaps zero -(-0.0439, 0.1972) - which suggests it is not really important, and was only found to be in the GLM/GAM analyses due to spatial autocorrelation in the response data.

#### 5.4.2 Farnes

Here the AIC selects distance to colony, distance to shore, summ\_front, spring\_front and ss\_wave. BIC picks only distance to shore. LRT agrees with AIC, but the GAM model chosen drops distance to colony, despite the four remaining variables displaying linear relationships. As can be seen in Figure A.4.2.4, summ\_front and ss\_wave have small positive relationships with preference, whereas spring\_front has a small negative one. The strongest relationship (and with p<0.001) is for distance to shore, very strongly negative, suggesting the birds like staying closer to shore.

There is no evidence in the credible intervals from the spatial INLA analysis that spatial autocorrelation influenced the model selection.

### 5.4.3 Forvie

For Forvie, the AIC selects distance to shore, sst\_june and strat\_temp. BIC drops sst\_june, but the LRT selection matches that of AIC. The GAM selection agrees with BIC, composed of distance to shore and strat\_temp. The effect of both of these variables is strongly negative (see Figure A.4.3.4). The spatial INLA analysis does not cast doubt on the earlier model selection.

### 5.4.4 Larne Lough

This data set proved very problematic; there were so few combinations of covariates for the case data, it was only feasible to use a very limited set of covariates. In fact, we chose distance to colony and distance to shore, as these were most likely to be important and were at least likely to be measured accurately. Both variables were indeed selected by *all* the different criteria. There was no evidence of the effects varying between years, and the spatial INLA analysis did not provide credible intervals overlapping zero.

From Figure A.4.4.4, it can be seen that the relationship between preference and both distance to colony and distance to shore is negative, and that the effect of distance to shore is stronger (with p=0.002).

### 5.5 Overall Comments

Looking across the colonies for each species, we find a number of consistent features. Note, however, that these are only associations; we cannot state that the selected covariates are causal drivers of foraging behaviour. Of course, it may be that the true drivers are in fact unrecorded.

For Arctic terns, distance to colony is an important predictor, and each analysis unearths one of the sst\_variables. Note that owing to correlations between the covariates, it is not possible to disentangle these effects further. Distance to colony is of course related to accessibility, but it is noticeable that most analyses find significant effects over and above this variable.

The situation for common terns would seem to be similar to that for Arctic terns, in that both distance to colony and an sst\_ variable are selected consistently. Mull is an exception here, but as noted earlier the geography surrounding that colony is problematic, in that is has likely caused biases in the data collected.

There is only one colony for roseate terns, which adds chl\_ variables to the distance to colony and sst\_selections from the previous two species.

For Sandwich terns, the only consistent explanatory variable chosen was distance to shore; the individual colonies mostly produced other significant effects, but these varied quite widely.

#### 6. Discussion

The methodology used presents an advance on earlier methods, with the weighted regression accounting for the differing numbers of observations in each of the recorded tracks. This method is different from that developed by Aarts *et al.* (2008), which does not explicitly use weights and hence implicitly ascribes more weight to tracks with more locations along them. We have made this alteration because the high correlation between locations on the same track makes track a better sampling unit than location

The maps in Section 7 illustrate clearly that the models describe some but not all of the predictability for preference of search/foraging area. With some maps – for example, the common tern for Coquet, the Sandwich tern for Forvie – the red shaded areas (representing higher preference) coincide well with the observed data. For others, however, the observed data lie on a mixture of red and blue areas, suggesting that the given explanatory variables are not sufficient to explain well the preferred foraging areas for those species/colonies; examples here include the common tern for Larne Lough or the Arctic tern for Outer Ards. However, note that we would expect to see some observations in the low preference areas as they are spatially more extensive.

Some of the large sample sizes effectively mean the analysis should be very sensitive to genuine relationships between tern presence and the environmental covariates. However, there is also a real possibility of uncovering spurious relationships, that is, relationships which may be due to a very small actual effect, caused by only a small number of observations that are outlying in covariate space. Further work should investigate the effect of removing outliers in the environmental covariates that may be unreliable. Despite this caveat, there was some evidence of consistent effects between colonies within species of tern. It may be possible that such differences within species but between colonies are due to the birds reacting to local variability. Future work could look at the extent to which it is possible to make reliable predictions for one colony using a model developed for another colony.

### 7. Maps Appendix

The maps in this appendix have been produced on the basis of the selected GAM models described in Section 5 and in the Results Appendix. Some outliers in the covariate values were removed before making predictions. For each species/colony combination, there are two pages of maps. The first page contains mapped preference, the second mapped usage. For ease of comparison with actual observations, each predicted map has been plotted twice – once by itself (top) and once with the observations added as black crosses.

For the usage maps, the probabilities have been multiplied by 1000.



## Arctic Tern, Coquet Colony – Preference



Arctic Tern, Coquet Colony – Usage









## Arctic Tern, Farnes Colony – Usage







## Arctic Tern, Outer Ards Colony – Usage







## **Common Tern, Coquet Colony – Usage**







### Common Tern, Larne Lough Colony – Usage











## **Common Tern, Mull Colony – Preference**



### Common Tern, Mull Colony – Usage


**Roseate Tern, Coquet Colony – Preference** 



**Roseate Tern, Coquet Colony – Usage** 



# Sandwich Tern, Coquet Colony – Preference



# Sandwich Tern, Coquet Colony – Usage



Sandwich Tern, Farnes Colony – Preference



# Sandwich Tern, Farnes Colony – Usage



# Sandwich Tern, Forvie Colony – Preference



# Sandwich Tern, Forvie Colony – Usage



Sandwich Tern, Larne Lough Colony – Preference





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Biomathematics and Statistics Scotland

# To carry out tern modelling under the Framework Agreement C10-0206-0387

CONTRACT No: C10-0206-0387

# **RESULTS APPENDIX**

Submitted to:

# **Joint Nature Conservation Committee**

# March 2012

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## A. Results

This appendix contains the text and graphical output from the analysis of the terns tracking data. The section match those in the results section of the main report, ordered by tern species and then by colony within species.

The sections within each species/colony sections are: (i) output from the AIC selection for GLM; (ii) output from the BIC selection for GLM; (iii) output from the likelihood ratio (LRT) selection for GLM, with checking for consistent Year effects if more than one year of data exists for that species/colony; (iv) output from the GAM selection, including plots of smooth covariate relationships; and (v) output from the INLA analysis applied to the selected GAM model, with plots of smooth covariates if linear terms were insufficient.

Proper significance assessment of GLMs is best performed by studying the Analysis of Deviance table, hence this is presented along with the standard "regression-style" table.

# A.1 Arctic Terns

# A.1.1 Coquet

## A.1.1.1 GLM Output using AIC for Selection

```
AIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_col + sst_may + summ_front_sd +
   ss_wave + ss_current, family = "binomial", data = complete.data.to.analyse,
   weights = weights)
Deviance Residuals:
    Min
                  Median
                                 30
              10
                                          Max
-0.96383 -0.03654 -0.01648 -0.00803 1.92561
Coefficients:
             Estimate Std. Error z value Pr(>|z|)
            -14.90185
                       7.39603 -2.015 0.0439 *
(Intercept)
             -0.20735 0.03635 -5.704 1.17e-08 ***
dist col
             1.53373 0.80628 1.902
                                        0.0571 .
sst may
summ front sd -0.02659 0.01193 -2.229 0.0258 *
             -0.03882
                        0.01675 -2.318
                                          0.0204 *
ss wave
             1.14923 0.54748 2.099 0.0358 *
ss_current
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
```

```
Null deviance: 542.78 on 47731 degrees of freedom
Residual deviance: 419.43 on 47726 degrees of freedom
AIC: 53.701
Number of Fisher Scoring iterations: 8
```

#### A.1.1.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_col, family = "binomial",
   data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q
                                        Max
-0.79352 -0.03705 -0.01715 -0.00702 1.93099
Coefficients:
           Estimate Std. Error z value Pr(>|z|)
                     0.18258 -4.914 8.91e-07 ***
(Intercept) -0.89725
dist col -0.21127 0.02803 -7.537 4.81e-14 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
   Null deviance: 542.78 on 47731 degrees of freedom
Residual deviance: 429.21 on 47730 degrees of freedom
```

```
AIC: 46.903
Number of Fisher Scoring iterations: 8
```

### A.1.1.3 GLM Output using Likelihood Ratio Tests for Selection

-0.96383 -0.03654 -0.01648 -0.00803 1.92561

Coefficients:

	Estimate	Std. Error	z value	Pr(> z )	
(Intercept)	-14.90185	7.39603	-2.015	0.0439	*
dist_col	-0.20735	0.03635	-5.704	1.17e-08	* * *
sst_may	1.53373	0.80628	1.902	0.0571	
summ_front_sd	-0.02659	0.01193	-2.229	0.0258	*
ss_wave	-0.03882	0.01675	-2.318	0.0204	*
ss_current	1.14923	0.54748	2.099	0.0358	*
Signif. codes:	0 ****	0.001 `**'	0.01 `*'	0.05 \.	0.1 1
(Dispersion pa	rameter fo	or binomial	family t	aken to k	pe 1)

Null deviance: 542.78 on 47731 degrees of freedom Residual deviance: 419.43 on 47726 degrees of freedom AIC: 53.701 Number of Fisher Scoring iterations: 8

#### Analysis of deviance output for reliable assessment of significance:

Now checking for interactions with Year:

#### A.1.1.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(dist_col, k = 3) + s(sst_may, k = 3) + s(summ_front_sd,
   k = 3) + s(ss_wave, k = 3) + s(ss_current, k = 3)
Parametric coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) -3.8379 0.2935 -13.08 <2e-16 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
                 edf Ref.df Chi.sq p-value
s(dist col)
              1.726 1.925 31.177 1.49e-07 ***
              1.000 1.000 3.664 0.0556 .
s(sst may)
s(summ_front_sd) 1.000 1.000 3.591 0.0581.
s(ss wave) 1.000 1.000 6.240 0.0125 *
s(ss current) 1.000 1.000 2.897 0.0887.
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

R-sq.(adj) = 0.196 Deviance explained = 23.5% REML score = 213.08 Scale est. = 1 n = 47732



Figure A.1.1.4: GAM covariate plots for Arctic Terns in Coquet

### A.1.1.5 INLA Output

Running INLA.

Call:

```
c("inla(formula = formula.inla, family = \"binomial\", data =
complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")
```

Fixed effects:

(Intercept)	-15.3423	7.3532	-30.1945	-15.1883	-1.3298	0.0022
sst_may	1.6314	0.8038	0.0994	1.6149	3.2528	0.0023
summ_front_sd	-0.0220	0.0121	-0.0465	-0.0217	0.0009	0.0014
ss_wave	-0.0451	0.0172	-0.0804	-0.0445	-0.0129	0.0017
ss_current	0.9403	0.5578	-0.1918	0.9532	1.9998	0.0027
dist_col1	-9.2197	1.7094	-12.9319	-9.0862	-6.2210	0.0837
dist_col2	-0.0356	3.1457	-7.1358	0.3443	5.1207	0.0013

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 7.547 3.169 1.368 7.527 13.816 Expected number of effective parameters(std dev): 7.001(0.001588) Number of equivalent replicates : 6818.26 Marginal Likelihood: -211.91



Figure A.1.1.5: INLA GAM covariate plots for Arctic Terns in Coquet (non-linear terms only)

# A.1.2 Farnes

## A.1.2.1 GLM Output using AIC for Selection

AIC Selected Model:

Call:

glm(formula = SEARCH\_FORAGE ~ dist\_col + sst\_april + summ\_front\_sd,

family = "binomial", data = complete.data.to.analyse, weights = weights)
Deviance Residuals:

Min 1Q Median 3Q Max -0.68797 -0.04791 -0.01884 -0.00604 1.58280

Coefficients:

Estimate Std. Error z value Pr(>|z|) (Intercept) 1.51958 1.12428 1.352 0.176505 dist\_col -0.30400 0.06477 -4.693 2.69e-06 \*\*\* sst\_april -0.47563 0.17774 -2.676 0.007451 \*\* summ\_front\_sd 0.04579 0.01281 3.574 0.000351 \*\*\* ----Signif. codes: 0 `\*\*\*' 0.001 `\*\*' 0.01 `\*' 0.05 `.' 0.1 ` ' 1 (Dispersion parameter for binomial family taken to be 1)

Null deviance: 213.73 on 17470 degrees of freedom Residual deviance: 159.82 on 17467 degrees of freedom AIC: 15.356 Number of Fisher Scoring iterations: 8

### A.1.2.2 GLM Output using BIC for Selection

```
summ_front_sd 0.03561 0.01171 3.040 0.002366 **
---
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
Null deviance: 213.73 on 17470 degrees of freedom
Residual deviance: 166.41 on 17468 degrees of freedom
AIC: 13.721
Number of Fisher Scoring iterations: 8
```

#### A.1.2.3 GLM Output using Likelihood Ratio Tests for Selection

```
LRT Selected Model:
Call:
glm(formula = formula.glm, family = "binomial", data = complete.data.to.analyse,
   weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q
                                        Max
-0.68797 -0.04791 -0.01884 -0.00604 1.58280
Coefficients:
           Estimate Std. Error z value Pr(>|z|)
            1.51958 1.12428 1.352 0.176505
(Intercept)
          -0.30400 0.06477 -4.693 2.69e-06 ***
dist col
sst_april -0.47563 0.17774 -2.676 0.007451 **
summ front sd 0.04579 0.01281 3.574 0.000351 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
```

Null deviance: 213.73 on 17470 degrees of freedom Residual deviance: 159.82 on 17467 degrees of freedom AIC: 15.356 Number of Fisher Scoring iterations: 8 Analysis of deviance output for reliable assessment of significance:

#### A.1.2.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH FORAGE ~ s(dist col, k = 3) + s(sst april, k = 3) + s(summ front sd,
   k = 3)
Parametric coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) -4.1271 0.5445 -7.579 3.48e-14 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
               edf Ref.df Chi.sq p-value
                1 1.000 22.02 2.69e-06 ***
s(dist col)
                1 1.001 7.16 0.007464 **
s(sst april)
s(summ front sd) 1 1.000 12.78 0.000351 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.357 Deviance explained = 25.2%
REML score = 82.13 Scale est. = 1 n = 17471
```



Figure A.1.2.4: GAM covariate plots for Arctic Terns in Farnes

## A.1.2.5 INLA Output

Running INLA.

Call:

c("inla(formula = formula.inla, family = \"binomial\", data = complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	1.5580	1.1243	-0.6075	1.5433	3.8060	0.0006
dist_col	-0.3095	0.0648	-0.4475	-0.3055	-0.1930	0.0036

sst_april	-0.4847 0.1777	-0.8406	-0.4822	-0.1428	0.0013
summ_front_sd	0.0471 0.0128	0.0229	0.0468	0.0732	0.0055

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 7.353 3.162 1.137 7.356 13.557 Expected number of effective parameters(std dev): 4.00(2.882e-05) Number of equivalent replicates : 4367.37 Marginal Likelihood: -87.00

## A.1.3 Outer Ards

### A.1.3.1 GLM Output using AIC for Selection

```
AIC Selected Model:
```

Call:

```
glm(formula = SEARCH_FORAGE ~ dist_col + dist_shore + chl_apr +
```

chl\_june + sst\_may, family = "binomial", data = complete.data.to.analyse,

weights = weights)

Deviance Residuals:

Min	1Q	Median	3Q	Max
-0.77316	-0.02755	-0.01621	-0.00902	0.69414

Coefficients:

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) 17.83795 7.34100 2.430 0.01510 *

dist_col -0.08642 0.06001 -1.440 0.14988

dist_shore -0.21399 0.13838 -1.546 0.12200

chl_apr -1.56145 0.53550 -2.916 0.00355 **

chl_june -1.07008 0.74078 -1.445 0.14859

sst_may -1.32740 0.49672 -2.672 0.00753 **

---

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1

(Dispersion parameter for binomial family taken to be 1)
```

Null deviance: 127.053 on 25374 degrees of freedom

Residual deviance: 94.462 on 25369 degrees of freedom AIC: 12.482 Number of Fisher Scoring iterations: 9

### A.1.3.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ chl_apr + sst_may, family = "binomial",
   data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
          1Q Median 3Q Max
    Min
-0.74926 -0.03023 -0.02159 -0.01476 0.88725
Coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) 5.0965
                     2.2975 2.218 0.026531 *
                      0.3386 -2.290 0.022005 *
chl apr
           -0.7756
sst may
           -0.7022
                     0.1881 -3.733 0.000189 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
   Null deviance: 127.05 on 25374 degrees of freedom
Residual deviance: 105.91 on 25372 degrees of freedom
AIC: 6.7872
```

Number of Fisher Scoring iterations: 7

#### A.1.3.3 GLM Output using Likelihood Ratio Tests for Selection

-0.76655 -0.02840 -0.01699 -0.01067 0.63367

Coefficients:

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) 7.80494 2.86999 2.720 0.006538 **

dist_col -0.12616 0.05107 -2.470 0.013500 *

chl_apr -1.20876 0.43496 -2.779 0.005453 **

sst_may -0.69251 0.20513 -3.376 0.000736 ***

---

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1

(Dispersion parameter for binomial family taken to be 1)
```

Null deviance: 127.053 on 25374 degrees of freedom Residual deviance: 98.209 on 25371 degrees of freedom AIC: 8.4015 Number of Fisher Scoring iterations: 7

#### Analysis of deviance output for reliable assessment of significance:

#### Now checking for interactions with Year:

Model: SEARCH\_FORAGE ~ Year \* (dist\_col + chl\_apr + sst\_may) Df Deviance AIC LRT Pr(>Chi) <none> 86.958 24.051 Year:dist\_col 2 87.638 20.730 0.67937 0.7120

```
Year:chl_apr 2 87.542 20.634 0.58306 0.7471
Year:sst_may 2 89.873 22.965 2.91445 0.2329
No significant Year interactions.
```

#### A.1.3.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(dist_col, k = 3) + s(chl_apr, k = 3) + s(sst_april,
   k = 3)
Parametric coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) -3.4964 0.4115 -8.496 <2e-16 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
           edf Ref.df Chi.sq p-value
s(dist_col) 1 1 6.504 0.010764 *
s(chl_apr) 1 1 7.885 0.004984 **
s(sst april) 1 1 11.524 0.000687 ***
____
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.725 Deviance explained = 22.7%
REML score = 50.408 Scale est. = 1 n = 25375
```



Figure A.1.3.4: GAM covariate plots for Arctic Terns in Outer Ards

## A.1.3.5 INLA Output

Running INLA.

Call:

c("inla(formula = formula.inla, family = \"binomial\", data = complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	8.1383	2.8296	3.0648	7.9583	14.1855	0.0120
dist_col	-0.1372	0.0510	-0.2441	-0.1348	-0.0437	0.0099

chl_apr	-1.3165 0.4372	-2.2435 -1.	2916 -0.5249	0.0206
sst april	-0.8058 0.2306	-1.3010 -0.	7901 -0.3955	0.0051

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 8.175 2.588 3.543 7.984 13.741 Expected number of effective parameters(std dev): 4.001(0.000108) Number of equivalent replicates : 6341.73 Marginal Likelihood: -52.33

## A.2 Common Terns

### A.2.1 Coquet

#### A.2.1.1 GLM Output using AIC for Selection

```
AIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_col + chl_june + sst_april +
    ss_wave, family = "binomial", data = complete.data.to.analyse,
    weights = weights)
Deviance Residuals:
```

Min	Min 1Q		ЗQ	Max
-0.93935	-0.04197	-0.01365	-0.00534	2.21575

```
Coefficients:
```

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) 2.77050 1.64251 1.687 0.09165 .

dist_col -0.23482 0.04176 -5.623 1.88e-08 ***

chl_june 0.48865 0.12892 3.790 0.00015 ***

sst_april -0.67033 0.21828 -3.071 0.00213 **

ss_wave -0.07805 0.02523 -3.093 0.00198 **

---

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1

(Dispersion parameter for binomial family taken to be 1)
```

Null deviance: 430.38 on 29969 degrees of freedom

Residual deviance: 319.16 on 29965 degrees of freedom AIC: 22.509 Number of Fisher Scoring iterations: 8

### A.2.1.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_col, family = "binomial",
    data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q Max
-0.75936 -0.04582 -0.01874 -0.00808 2.24320
Coefficients:
        Estimate Std. Error z value Pr(>|z|)
(Intercept) -0.92484 0.20783 -4.450 8.59e-06 ***
dist_col -0.22190 0.03366 -6.592 4.33e-11 ***
---
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 `' 1
(Dispersion parameter for binomial family taken to be 1)
Null deviance: 430.38 on 29969 degrees of freedom
```

Residual deviance: 340.30 on 29968 degrees of freedom AIC: 16.789 Number of Fisher Scoring iterations: 8

### A.2.1.3 GLM Output using Likelihood Ratio Tests for Selection

```
LRT Selected Model:
Call:
glm(formula = formula.glm, family = "binomial", data = complete.data.to.analyse,
   weights = weights)
Deviance Residuals:
   Min      1Q   Median      3Q   Max
-0.93935 -0.04197 -0.01365 -0.00534     2.21575
```

Coefficients:

Estimate Std. Error z value Pr(>|z|) (Intercept) 2.77050 1.64251 1.687 0.09165. dist\_col -0.23482 0.04176 -5.623 1.88e-08 \*\*\* chl\_june 0.48865 0.12892 3.790 0.00015 \*\*\* sst\_april -0.67033 0.21828 -3.071 0.00213 \*\* ss\_wave -0.07805 0.02523 -3.093 0.00198 \*\* ----Signif. codes: 0 `\*\*\*' 0.001 `\*\*' 0.01 `\*' 0.05 `.' 0.1 ` ' 1 (Dispersion parameter for binomial family taken to be 1)

Null deviance: 430.38 on 29969 degrees of freedom Residual deviance: 319.16 on 29965 degrees of freedom AIC: 22.509 Number of Fisher Scoring iterations: 8

#### Analysis of deviance output for reliable assessment of significance:

#### Now checking for interactions with Year:

```
Model:
SEARCH_FORAGE ~ Year * (dist_col + chl_june + sst_april + ss_wave)
Df Deviance AIC LRT Pr(>Chi)
<none> 308.41 40.566
```

```
Year:dist_col 2 313.61 41.762 5.1958 0.07443 .
Year:chl_june 2 308.92 37.076 0.5104 0.77476
Year:sst_april 2 308.77 36.921 0.3549 0.83738
Year:ss_wave 2 311.43 39.589 3.0232 0.22056
---
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Year effects not consistent - run one year at a time.
```

### A.2.1.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(dist_col, k = 3) + s(chl_june, k = 3) + s(sst_april,
    k = 3) + s(ss_wave, k = 3)
Parametric coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) -5.1675 0.9453 -5.467 4.59e-08 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
              edf Ref.df Chi.sq p-value
s(dist col) 1.629 1.862 17.739 0.000115 ***
s(chl june) 1.000 1.000 12.278 0.000458 ***
s(sst april) 1.453 1.701 9.607 0.005681 **
s(ss wave) 1.000 1.000 5.993 0.014362 *
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.2 Deviance explained = 26.4%
REML score = 163.61 Scale est. = 1 n = 29970
```



Figure A.2.1.4: GAM covariate plots for Common Terns in Coquet

## A.2.1.5 INLA Output

Running INLA.

Call:

c("inla(formula = formula.inla, family = \"binomial\", data = complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	3.1326	1.9602	-0.3413	2.9926	7.3585	0.0000
chl_june	0.4647	0.1307	0.2149	0.4622	0.7280	0.0024

ss_wave	-0.0570	0.0270	-0.1140	-0.0555	-0.0077	0.0004
dist_col1	-16.7128	5.4688	-28.9271	-16.1196	-7.5617	0.0502
dist_col2	-20.8654	10.6971	-44.6611	-19.7513	-2.8460	0.0466
sst_april1	-10.1684	3.8209	-18.4069	-9.8953	-3.4036	0.0001
sst_april2	-0.5175	1.5212	-3.5177	-0.5112	2.4481	0.0004

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 7.344 3.162 1.146 7.343 13.558 Expected number of effective parameters(std dev): 6.84(5.269e-05) Number of equivalent replicates : 4381.76 Marginal Likelihood: -155.88



Figure A.2.1.5: INLA GAM covariate plots for Common Terns in Coquet (non-linear terms only)

# A.2.2 Larne Lough

### A.2.2.1 GLM Output using AIC for Selection

```
AIC Selected Model:
```

Call:

```
glm(formula = SEARCH_FORAGE ~ dist_col + dist_shore + sst_april +
```

bathy 1sec, family = "binomial", data = complete.data.to.analyse,

weights = weights)

Deviance Residuals:

Min	1Q	Median	3Q	Max
-0.83858	-0.01706	-0.00799	-0.00033	0.81678

Coefficients:

	Estimate	Std. Er	ror z	value	Pr(>	z )				
(Intercept)	0.69907	2.44	725	0.286	0.7	7751				
dist_col	-1.02236	0.39	887 -	-2.563	0.0	0104	*			
dist_shore	1.01402	0.42	836	2.367	0.0	)179	*			
sst_april	0.54383	0.28	704	1.895	0.0	)581	•			
bathy_1sec	0.04195	0.01	023	4.101	4.120	e-05	* * *			
Signif. code	es: 0 `**	** 0.00	1 \**'	0.01	<b>`*'</b> (	0.05	۰.′	0.1	`	<b>'</b> 1
(Dispersion	parameter	for bi	nomial	. famil	y tał	ken t	o be	e 1)		

Null deviance: 109.968 on 18056 degrees of freedom Residual deviance: 56.951 on 18052 degrees of freedom AIC: 10.953 Number of Fisher Scoring iterations: 10

#### A.2.2.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_col + dist_shore + bathy_lsec,
    family = "binomial", data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
```

Min 1Q Median 3Q Max

-1.06182 -0.01760 -0.00960 -0.00083 0.83903

Coefficients:

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) 2.85919 1.39032 2.056 0.03973 *

dist_col -0.81257 0.34627 -2.347 0.01894 *

dist_shore 0.91903 0.37410 2.457 0.01402 *

bathy_lsec 0.04069 0.01073 3.792 0.00015 ***

----

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1

(Dispersion parameter for binomial family taken to be 1)

Null deviance: 109.968 on 18056 degrees of freedom
```

Residual deviance: 66.235 on 18053 degrees of freedom AIC: 10.43 Number of Fisher Scoring iterations: 10

#### A.2.2.3 GLM Output using Likelihood Ratio Tests for Selection

```
bathy_1sec 0.04195 0.01023 4.101 4.12e-05 ***
```

sst april 0.54383 0.28704 1.895 0.0581.

---

```
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
Null deviance: 109.968 on 18056 degrees of freedom
Residual deviance: 56.951 on 18052 degrees of freedom
AIC: 10.953
Number of Fisher Scoring iterations: 10
```

#### Analysis of deviance output for reliable assessment of significance:

#### Now checking for interactions with Year:

#### A.2.2.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
```
```
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(dist_col, k = 3) + s(dist_shore, k = 3) + s(sst_april,
   k = 3) + s(bathy_1sec, k = 3)
Parametric coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) -7.601 1.677 -4.533 5.82e-06 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
              edf Ref.df Chi.sq p-value
s(dist_col) 1.000 1.000 7.239 0.00713 **
s(dist shore) 1.000 1.000 4.060 0.04391 *
s(sst april) 1.000 1.000 4.744 0.02940 *
s(bathy_1sec) 1.875 1.984 18.665 8.65e-05 ***
____
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.641 Deviance explained = 54.2%
REML score = 24.223 Scale est. = 1 n = 18057
```



Figure A.2.2.4: GAM covariate plots for Common Terns in Larne Lough

# A.2.2.5 INLA Output

Running INLA.

Call:

```
c("inla(formula = formula.inla, family = \"binomial\", data =
complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")
```

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	11.0152	8.6669	-7.0725	11.3943	26.9988	0.0034
dist_col	-1.0403	0.3609	-1.8549	-0.9966	-0.4494	0.0214

dist_shore	0.8705	0.3987	0.2148	0.8233	1.7681	0.0090
sst_april	0.6875	0.2883	0.2161	0.6523	1.3387	0.0239
bathy_1sec1	-27.9916	15.4704	-56.6066	-28.6403	4.2129	0.0003
bathy_1sec2	1.6795	3.0862	-3.8960	1.5006	8.2339	0.0104

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 7.356 3.172 1.171 7.341 13.611 Expected number of effective parameters(std dev): 5.751(0.0006167) Number of equivalent replicates : 3139.91 Marginal Likelihood: -23.28



Figure A.2.2.5: INLA GAM covariate plots for Common Terns in Larne Lough (non-linear terms only)

# A.2.3 Leith

## A.2.3.1 GLM Output using AIC for Selection

AIC Selected Model:

Call:

glm(formula = SEARCH\_FORAGE ~ dist\_col + dist\_shore + chl\_may +

chl\_june + sst\_may + summ\_front + spring\_front + sal\_spring +

bathy\_lsec + slope\_ls\_deg, family = "binomial", data = complete.data.to.analyse,

weights = weights)

Deviance Residuals:

Min	1Q	Median	ЗQ	Max
-0.85999	-0.04498	-0.02034	-0.00544	1.98457

Coefficients:

	Estimate	Std. Error	z value	Pr(> z )		
(Intercept)	-5.847e+02	2.060e+02	-2.838	0.00454	* *	
dist_col	-2.112e-01	4.369e-02	-4.833	1.35e-06	* * *	
dist_shore	-3.564e-01	1.362e-01	-2.618	0.00885	* *	
chl_may	1.721e-02	8.495e-03	2.026	0.04281	*	
chl_june	2.571e-02	1.212e-02	2.122	0.03386	*	
sst_may	1.146e-01	5.737e-02	1.998	0.04576	*	
summ_front	6.155e-02	3.372e-02	1.825	0.06793	•	
spring_front	-1.292e-01	8.593e-02	-1.504	0.13257		
sal_spring	1.661e+01	5.880e+00	2.824	0.00474	* *	
bathy_1sec	4.001e-02	1.830e-02	2.186	0.02880	*	
<pre>slope_1s_deg</pre>	1.557e-01	8.391e-02	1.856	0.06344	•	
Signif. codes	5: 0 `***'	0.001 `**'	0.01 `*'	0.05 `.'	0.1 1	
(Dispersion parameter for binomial family taken to be 1)						

1

Null deviance: 726.30 on 57616 degrees of freedom Residual deviance: 591.83 on 57606 degrees of freedom AIC: 74.249

Number of Fisher Scoring iterations: 8

#### A.2.3.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_col, family = "binomial",
   data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
    Min
             1Q Median 3Q
                                          Max
-0.76723 -0.04666 -0.02832 -0.01118 2.10430
Coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) -0.87617 0.18912 -4.633 3.61e-06 ***
dist col -0.22386 0.03131 -7.149 8.72e-13 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
   Null deviance: 726.30 on 57616 degrees of freedom
Residual deviance: 629.43 on 57615 degrees of freedom
AIC: 61.737
Number of Fisher Scoring iterations: 7
```

### A.2.3.3 GLM Output using Likelihood Ratio Tests for Selection

```
LRT Selected Model:
Call:
glm(formula = formula.glm, family = "binomial", data = complete.data.to.analyse,
   weights = weights)
Deviance Residuals:
   Min 1Q Median 3Q Max
-0.80966 -0.04530 -0.02140 -0.00603 2.02993
Coefficients:
```

```
Estimate Std. Error z value Pr(>|z|)
(Intercept) -5.271e+02 2.018e+02 -2.612 0.00900 **
```

```
dist_col -2.132e-01 4.310e-02 -4.947 7.52e-07 ***
dist_shore -3.141e-01 1.339e-01 -2.346 0.01899 *
chl_may 1.570e-02 8.337e-03 1.884 0.05962 .
chl_june 2.208e-02 1.157e-02 1.909 0.05627 .
sst_may 1.171e-01 5.700e-02 2.055 0.03988 *
sal_spring 1.496e+01 5.759e+00 2.599 0.00936 **
bathy_lsec 3.496e-02 1.768e-02 1.977 0.04802 *
slope_ls_deg 1.621e-01 8.369e-02 1.937 0.05278 .
----
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
```

Null deviance: 726.30 on 57616 degrees of freedom Residual deviance: 596.51 on 57608 degrees of freedom AIC: 70.309 Number of Fisher Scoring iterations: 8

#### Analysis of deviance output for reliable assessment of significance:

# Model: SEARCH\_FORAGE ~ dist\_col + dist\_shore + chl\_may + chl\_june + sst\_may + sal\_spring + bathy\_lsec + slope\_ls\_deg Df Deviance AIC LRT Pr(>Chi) 596.51 70.309 <none> dist col 1 640.09 111.887 43.577 4.075e-11 \*\*\* dist shore 1 602.38 74.172 5.862 0.015467 \* 1 600.08 71.874 3.564 0.059032 . chl\_may chl\_june 1 599.98 71.771 3.462 0.062798 . sst\_may 1 600.86 72.652 4.343 0.037161 \* sal\_spring 1 603.67 75.470 7.161 0.007451 \*\* bathy 1sec 1 600.85 72.648 4.339 0.037258 \* slope\_1s\_deg 1 600.07 71.869 3.560 0.059180 . \_\_\_\_ Signif. codes: 0 `\*\*\*' 0.001 `\*\*' 0.01 `\*' 0.05 `.' 0.1 ` ' 1

Now checking for interactions with Year:

```
Model:
SEARCH FORAGE ~ Year * (dist col + dist shore + chl may + chl june +
   sst may + sal spring + bathy 1sec + slope 1s deg)
                Df Deviance AIC LRT Pr(>Chi)
                     575.93 86.618
<none>
Year:dist_col
                1 576.17 84.866 0.2485 0.6181
Year:dist_shore 1 578.43 87.124 2.5061 0.1134
Year:chl_may 1 577.60 86.291 1.6729 0.1959
Year:chl_june
                1 576.83 85.518 0.9000 0.3428
Year:sst may
                1 579.45 88.144 3.5264 0.0604 .
Year:sal spring 1 576.19 84.879 0.2607 0.6096
Year:bathy_1sec 1 576.69 85.385 0.7670 0.3812
Year:slope 1s deg 1 576.18 84.874 0.2562 0.6128
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Year effects not consistent - run one year at a time.
```

## A.2.3.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(dist_col, k = 3) + s(dist_shore, k = 3) + s(chl_may,
        k = 3) + s(chl_june, k = 3) + s(sst_may, k = 3) + s(sal_spring,
        k = 3) + s(bathy_lsec, k = 3) + s(slope_ls_deg, k = 3)
Parametric coefficients:
            Estimate Std. Error z value Pr(>|z|)
(Intercept) -4.4153 0.6363 -6.939 3.96e-12 ***
---
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

```
edf Ref.df Chi.sq p-value
```

s(dist_col)	1.604	1.843	21.652	1.56e-05	* * *	
s(dist_shore)	1.000	1.000	6.131	0.0133	*	
s(chl_may)	1.000	1.000	3.534	0.0601		
s(chl_june)	1.000	1.000	3.262	0.0709		
s(sst_may)	1.000	1.000	3.376	0.0662		
s(sal_spring)	1.000	1.000	5.481	0.0192	*	
s(bathy_1sec)	1.511	1.761	4.944	0.0674		
s(slope_1s_deg)	1.004	1.009	3.288	0.0707		
Signif. codes:	0 '***'	0.001	<b>`**′</b> 0	.01 `*' 0	).05 <b>`.'</b> 0.1	<b>`′</b> 1
R-sq.(adj) = 0	.199 I	Devianc	e expla	ined = $18$	3.3%	
REML score = 3	07.6 Sc	cale es	t. = 1	r	n = 57617	



Figure A.2.3.4: GAM covariate plots for Common Terns in Leith

# A.2.3.5 INLA Output

Running INLA.

Call:

```
c("inla(formula = formula.inla, family = \"binomial\", data =
complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")
```

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	-15.5270	32.2965	-78.8599	-15.5262	47.8138	0.0000
dist_shore	-0.3160	0.1331	-0.5826	-0.3142	-0.0600	0.0004

chl_may	0.0153	0.0082	-0.0007	0.0153	0.0314	0.0011
chl_june	0.0246	0.0116	0.0015	0.0247	0.0471	0.0000
sst_may	0.1286	0.0550	0.0231	0.1278	0.2389	0.0001
sal_spring	0.1967	0.9223	-1.6119	0.1966	2.0057	0.0002
slope_1s_deg	0.1728	0.0857	0.0066	0.1721	0.3429	0.0009
dist_col1	-13.3207	4.7528	-23.8067	-12.8663	-5.2110	0.0024
dist_col2	-20.3520	10.6448	-43.6668	-19.4126	-1.9789	0.0024
bathy_1sec1	9.5332	4.6518	1.2654	9.2101	19.5414	0.0056
bathy_1sec2	1.0313	0.8747	-0.6310	1.0118	2.8038	0.0076

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 7.159 3.001 1.194 7.195 12.961 Expected number of effective parameters(std dev): 9.865(0.0262) Number of equivalent replicates : 5840.40 Marginal Likelihood: -305.98



Figure A.2.3.5: INLA GAM covariate plots for Common Terns in Leith (non-linear terms only)

# A.2.4 Mull

### A.2.4.1 GLM Output using AIC for Selection

```
AIC Selected Model:
```

```
Call:
```

```
glm(formula = SEARCH_FORAGE ~ chl_apr + chl_may + sst_april +
```

sst\_may + ss\_wave, family = "binomial", data = complete.data.to.analyse,

1

weights = weights)

Deviance Residuals:

Min	1Q	Median	ЗQ	Max
-0.83221	-0.03256	-0.00769	-0.00056	1.41323

Coefficients:

	Estimate	Std. Error	z value	Pr(> z )	
(Intercept)	-5.12136	1.76890	-2.895	0.00379	* *
chl_apr	0.05560	0.02499	2.225	0.02609	*
chl_may	0.02995	0.01629	1.839	0.06597	•
sst_april	-0.25411	0.18224	-1.394	0.16320	
sst_may	0.41777	0.27283	1.531	0.12572	
ss_wave	-3.43918	2.66105	-1.292	0.19621	
Signif. code	es: 0 `**	**′ 0.001 ``	**′ 0.01	<b>`*′</b> 0.05	`.' 0.1 `'
(Dispersion	parameter	for binom:	ial famil	Ly taken t	to be 1)

Null deviance: 91.876 on 7415 degrees of freedom Residual deviance: 59.726 on 7410 degrees of freedom AIC: 15.649 Number of Fisher Scoring iterations: 11

### A.2.4.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ chl_apr, family = "binomial", data =
complete.data.to.analyse,
    weights = weights)
```

Deviance Residuals:

Min	1Q	Median	3Q	Max		
-0.69210 -0.	.03973 -	0.02633	-0.01780	1.75873		
Coefficients	:					
I	Estimate	Std. Errc	or z value	Pr(> z )		
(Intercept) -	-4.15516	0.5857	9 -7.093	1.31e-12	* * *	
chl_apr	0.06450	0.0154	3 4.180	2.92e-05	* * *	
Signif. codes	s: 0 `**	*′ 0.001	<b>`**'</b> 0.01	<b>`*'</b> 0.05	`.' 0.1	<b>`′</b> 1
(Dispersion p	parameter	for bind	mial fami	ly taken t	to be 1)	
Null devi	iance: 91	.876 on	7415 deg	rees of fi	reedom	
Residual devi	iance: 72	.609 on	7414 deg	rees of fi	reedom	
AIC: 8.593						

Number of Fisher Scoring iterations: 7

### A.2.4.3 GLM Output using Likelihood Ratio Tests for Selection

```
LRT Selected Model:
Call:
glm(formula = formula.glm, family = "binomial", data = complete.data.to.analyse,
   weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q Max
-0.87668 -0.03694 -0.01094 -0.00133 1.46745
Coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) -3.28231 0.82000 -4.003 6.26e-05 ***
chl_apr 0.03785 0.01768 2.141 0.0323 *
chl may
          0.02483 0.01421 1.748 0.0805 .
ss wave -2.71678 2.06153 -1.318 0.1876
____
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

```
(Dispersion parameter for binomial family taken to be 1)
Null deviance: 91.876 on 7415 degrees of freedom
Residual deviance: 62.608 on 7412 degrees of freedom
AIC: 11.897
Number of Fisher Scoring iterations: 11
```

### Analysis of deviance output for reliable assessment of significance:

## A.2.4.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(chl_apr, k = 3) + s(chl_may, k = 3) + s(bathy_lsec,
        k = 3)
Parametric coefficients:
        Estimate Std. Error z value Pr(>|z|)
(Intercept) -4.5358 0.7937 -5.714 1.1e-08 ***
---
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
        edf Ref.df Chi.sq p-value
s(chl_apr) 1.741 1.927 11.552 0.00283 **
```

```
s(chl_may) 1.613 1.845 5.461 0.05614 .
s(bathy_1sec) 1.799 1.958 8.269 0.01528 *
---
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.329 Deviance explained = 40.1%
REML score = 31.679 Scale est. = 1 n = 7416
```



Figure A.2.4.4: GAM covariate plots for Common Terns in Mull

bathy\_1sec

### A.2.4.5 INLA Output

Running INLA.

Call:

c("inla(formula = formula.inla, family = \"binomial\", data = complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	-2.1257	1.6735	-5.7285	-2.0083	0.8434	0.0053
chl_apr1	10.8786	4.2873	3.3763	10.5368	20.1982	0.0213
chl_apr2	1.5918	1.8559	-1.9418	1.5528	5.3451	0.0039
chl_may1	5.3970	2.5115	0.8383	5.2607	10.7080	0.0216
chl_may2	1.6592	1.0339	-0.3856	1.6654	3.6693	0.0163
bathy_1sec1	-12.7551	4.2179	-21.6618	-12.5304	-5.0780	0.0420
bathy_1sec2	-0.2244	1.6295	-3.5639	-0.1747	2.8387	0.0001

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 7.1401 3.1617 0.9245 7.1389 13.3516 Expected number of effective parameters(std dev): 6.95(1.466e-05) Number of equivalent replicates : 1066.98 Marginal Likelihood: -16.99



Figure A.2.4.5: INLA GAM covariate plots for Common Terns in Mull (non-linear terms only)

# A.3 Roseate Terns

# A.3.1 Coquet

## A.3.1.1 GLM Output using AIC for Selection

```
AIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_col + dist_shore + chl_may +
    sst_june + sst_may + summ_front_sd + strat_temp + ss_current,
    family = "binomial", data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
```

Min	1Q	Median	3Q	Max
-1.01134	-0.01394	-0.00392	-0.00076	1.25435

Coefficients:								
	Estimate Sto	d. Error z	value	Pr(> z )				
(Intercept)	-6.25604	8.35046	-0.749	0.45375				
dist_col	-0.38296	0.12664	-3.024	0.00249	* *			
dist_shore	-0.99891	0.47034	-2.124	0.03369	*			
chl_may	-0.13153	0.08050	-1.634	0.10229				
sst_june	-5.83586	1.39780	-4.175	2.98e-05	* * *			
sst_may	6.28284	1.52123	4.130	3.63e-05	* * *			
summ_front_sd	0.06155	0.03089	1.993	0.04632	*			
strat_temp	6.06231	2.92517	2.072	0.03822	*			
ss_current	-2.33588	1.67821	-1.392	0.16396				
Signif. codes:	: 0 `***′ 0	.001 `**′	0.01 `*	· 0.05 `.	.′ 0.1 `′	1		
(Dispersion pa	(Dispersion parameter for binomial family taken to be 1)							
Null devia	ance: 207.04	on 24099	) degre	es of fre	eedom			
Residual devia	ance: 89.69	on 24091	degre	es of fre	eedom			

AIC: 25.75

Number of Fisher Scoring iterations: 10

# A.3.1.2 GLM Output using BIC for Selection

BIC Selected Model:

#### Call:

-1.09199 -0.01590 -0.00607 -0.00152 1.28785

Coefficients:

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) 2.36373 3.27227 0.722 0.47008

dist_col -0.23490 0.07924 -2.964 0.00303 **

sst_june -4.10827 0.73106 -5.620 1.91e-08 ***

sst_may 4.75048 0.87728 5.415 6.13e-08 ***

----

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1

(Dispersion parameter for binomial family taken to be 1)

Null deviance: 207.04 on 24099 degrees of freedom

Residual deviance: 102.27 on 24096 degrees of freedom

AIC: 18.038

Number of Fisher Scoring iterations: 9
```

### A.3.1.3 GLM Output using Likelihood Ratio Tests for Selection

```
LRT Selected Model:
Call:
glm(formula = formula.glm, family = "binomial", data = complete.data.to.analyse,
  weights = weights)
Deviance Residuals:
    Min         1Q   Median          3Q          Max
-1.18335  -0.01465  -0.00634  -0.00280          1.19303
```

```
Coefficients:
```

	Estimate	Std. Error	z value	Pr(> z )	
(Intercept)	8.53713	4.55846	1.873	0.06109	
dist_col	-0.23031	0.08506	-2.707	0.00678	* *
chl_may	-0.21875	0.09195	-2.379	0.01736	*
sst_june	-6.00578	1.33205	-4.509	6.52e-06	* * *
sst_may	6.43205	1.40897	4.565	4.99e-06	* * *
summ_front_sd	0.05353	0.02624	2.040	0.04132	*
chl_apr	0.30112	0.15868	1.898	0.05775	
Signif. codes	: 0 `***'	0.001 `**'	0.01 `	*′ 0.05 <b>`</b> .	.′ 0.1 `′ 1

(Dispersion parameter for binomial family taken to be 1)

Null deviance: 207.045 on 24099 degrees of freedom Residual deviance: 91.709 on 24093 degrees of freedom AIC: 22.587 Number of Fisher Scoring iterations: 9

```
Analysis of deviance output for reliable assessment of significance:
```

```
Model:
SEARCH_FORAGE ~ dist_col + chl_may + sst_june + sst_may + summ_front_sd +
```

```
Df Deviance AIC LRT Pr(>Chi)

<none> 91.709 22.587

dist_col 1 102.701 31.579 10.991 0.0009154 ***

chl_may 1 99.205 28.083 7.495 0.0061857 **

sst_june 1 123.777 52.655 32.068 1.489e-08 ***

sst_may 1 128.140 57.018 36.431 1.582e-09 ***

summ_front_sd 1 95.438 24.316 3.729 0.0534870 .

chl_apr 1 95.190 24.068 3.481 0.0620895 .

----
```

#### Now checking for interactions with Year:

#### Model:

```
SEARCH_FORAGE ~ Year * (dist_col + chl_may + sst_june + sst_may +
summ front sd + chl apr)
```

 Df Deviance
 AIC
 LRT Pr(>Chi)

 <none>
 78.172
 49.101

 Year:dist\_col
 2
 78.598
 45.527
 0.4260
 0.8082

 Year:chl\_may
 2
 80.133
 47.062
 1.9605
 0.3752

 Year:sst\_june
 2
 81.030
 47.959
 2.8577
 0.2396

 Year:sst\_may
 2
 82.214
 49.143
 4.0415
 0.1326

 Year:summ\_front\_sd
 2
 80.344
 47.273
 2.1719
 0.3376

 Year:chl\_apr
 2
 78.421
 45.350
 0.2487
 0.8831

No significant Year interactions.

### A.3.1.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(dist_col, k = 3) + s(chl_apr, k = 3) + s(chl_may,
   k = 3) + s(sst_june, k = 3) + s(sst_may, k = 3) + s(summ_front_sd,
   k = 3)
Parametric coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) -5.6329 0.8417 -6.692 2.2e-11 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
              edf Ref.df Chi.sq p-value
s(dist_col) 1 1 7.331 0.00678 **
s(chl_apr) 1 1 3.601 0.05775.
               1 1 5.659 0.01737 *
s(chl_may)
                1
                      1 20.328 6.52e-06 ***
s(sst_june)
               1 1 20.840 4.99e-06 ***
s(sst_may)
s(summ front sd) 1 1 4.163 0.04132 *
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.766 Deviance explained = 55.7\%
REML score = 46.383 Scale est. = 1 n = 24100
```



Figure A.3.1.4: GAM covariate plots for Roseate Terns in Coquet

# A.3.1.5 INLA Output

Running INLA.

Call:

c("inla(formula = formula.inla, family = \"binomial\", data = complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	7.8512	4.4905	-1.4586	8.0255	16.1874	0.0037
dist_col	-0.2359	0.0845	-0.4180	-0.2299	-0.0860	0.0036

chl_apr	0.3065	0.1579	-0.0034	0.3066	0.6158	0.0008
chl_may	-0.2116	0.0902	-0.4015	-0.2070	-0.0469	0.0000
sst_june	-5.9285	1.2768	-8.6319	-5.8577	-3.6127	0.0044
sst_may	6.5034	1.3495	4.0569	6.4286	9.3604	0.0221
summ_front_sd	0.0535	0.0258	0.0013	0.0540	0.1028	0.0047

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 7.424 0.340 6.756 7.424 8.092 Expected number of effective parameters(std dev): 6.969(0.001662) Number of equivalent replicates : 3458.37 Marginal Likelihood: -55.62

# A.4 Sandwich Terns

## A.4.1 Coquet

### A.4.1.1 GLM Output using AIC for Selection

```
AIC Selected Model:
Call:
glm(formula = SEARCH FORAGE ~ dist col + dist shore + chl june +
   sst april, family = "binomial", data = complete.data.to.analyse,
   weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q
                                       Max
-1.01383 -0.01093 -0.00155 -0.00016 2.28876
Coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) 1.75429 0.77899 2.252 0.024322 *
dist col -0.09850 0.02570 -3.832 0.000127 ***
dist_shore -0.25459 0.09668 -2.633 0.008459 **
          0.10420 0.05848 1.782 0.074752 .
chl june
sst april -0.38100 0.11822 -3.223 0.001270 **
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

(Dispersion parameter for binomial family taken to be 1) Null deviance: 447.26 on 55926 degrees of freedom Residual deviance: 261.95 on 55922 degrees of freedom AIC: 55.537 Number of Fisher Scoring iterations: 10

### A.4.1.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH FORAGE ~ dist col + dist shore, family = "binomial",
   data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q Max
-0.93590 -0.00920 -0.00064 -0.00003 2.15522
Coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) 0.19595 0.29277 0.669 0.503311
dist_col -0.08246 0.02221 -3.712 0.000206 ***
dist shore -0.44069 0.09469 -4.654 3.26e-06 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
   Null deviance: 447.26 on 55926 degrees of freedom
Residual deviance: 274.44 on 55924 degrees of freedom
AIC: 57.509
```

### A.4.1.3 GLM Output using Likelihood Ratio Tests for Selection

Number of Fisher Scoring iterations: 10

```
LRT Selected Model:
Call:
glm(formula = formula.glm, family = "binomial", data = complete.data.to.analyse,
```

```
weights = weights)
```

Deviance Residuals:

Min	1Q	Median	3Q	Max
-1.01383	-0.01093	-0.00155	-0.00016	2.28876

Coefficients:

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) 1.75429 0.77899 2.252 0.024322 *

dist_col -0.09850 0.02570 -3.832 0.000127 ***

dist_shore -0.25459 0.09668 -2.633 0.008459 **

chl_june 0.10420 0.05848 1.782 0.074752 .

sst_april -0.38100 0.11822 -3.223 0.001270 **

----

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

(Dispersion parameter for binomial family taken to be 1)

Null deviance: 447.26 on 55926 degrees of freedom Residual deviance: 261.95 on 55922 degrees of freedom AIC: 55.537 Number of Fisher Scoring iterations: 10

#### Analysis of deviance output for reliable assessment of significance:

Now checking for interactions with Year:

#### A.4.1.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(dist_col, k = 3) + s(dist_shore, k = 3) + s(chl june,
    k = 3) + s(sst_april, k = 3)
Parametric coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) -13.913 4.196 -3.315 0.000915 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
              edf Ref.df Chi.sq p-value
s(dist col) 1.804 1.959 7.345 0.02434 *
s(dist shore) 1.000 1.000 8.042 0.00457 **
s(chl june) 1.000 1.000 1.530 0.21619
s(sst april) 1.736 1.928 10.991 0.00376 **
```

Signif. codes: 0 `\*\*\*' 0.001 `\*\*' 0.01 `\*' 0.05 `.' 0.1 ` ' 1

R-sq.(adj) = 0.461 Deviance explained = 42.9% REML score = 132.43 Scale est. = 1 n = 55927



Figure A.4.1.4: GAM covariate plots for Sandwich Terns in Coquet

### A.4.1.5 INLA Output

Running INLA.

Call:

```
c("inla(formula = formula.inla, family = \"binomial\", data =
complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")
```

Fixed effects:

---

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	0.4786	0.8183	-1.1047	0.4706	2.1073	0.0028
dist_shore	-0.2600	0.0989	-0.4706	-0.2542	-0.0808	0.0030
chl_june	0.0771	0.0614	-0.0439	0.0773	0.1972	0.0022
dist_col1	-20.7704	8.5412	-37.6322	-20.6781	-4.0885	0.0128
dist_col2	-33.6238	16.2892	-65.5195	-33.5422	-1.7116	0.0154
sst_april1	-1.4888	1.6015	-4.6848	-1.4706	1.6020	0.0002
sst_april2	-3.2350	1.2110	-5.6804	-3.2115	-0.9226	0.0016

Model hyperparameters:

			mean	sd	0.025quant	0.5quant	0.975quant
т.0	for	mesh.points-basisT	8.034	3.144	1.830	8.052	14.162

Expected number of effective parameters(std dev): 6.661(4.235e-05) Number of equivalent replicates : 8395.72 Marginal Likelihood: -126.07



Figure A.4.1.5: INLA GAM covariate plots for Sandwich Terns in Coquet (non-linear terms only)

# A.4.2 Farnes

## A.4.2.1 GLM Output using AIC for Selection

-0.5551 0.0000 0.0000 0.0000 1.3428

Coefficients:

\_\_\_

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) 1.68802 1.67480 1.008 0.31351

dist_col -0.16160 0.11039 -1.464 0.14321

dist_shore -3.51590 0.97701 -3.599 0.00032 ***

summ_front 0.28026 0.11436 2.451 0.01426 *

spring_front -0.46366 0.19019 -2.438 0.01477 *

ss_wave 0.04123 0.01939 2.126 0.03353 *

---

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1

(Dispersion parameter for binomial family taken to be 1)
```

Null deviance: 160.42 on 15038 degrees of freedom Residual deviance: 30.82 on 15033 degrees of freedom AIC: 12.472

Number of Fisher Scoring iterations: 14

### A.4.2.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_shore, family = "binomial",
    data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q Max
-0.64514 -0.00004 0.00000 0.00000 1.63006
Coefficients:
        Estimate Std. Error z value Pr(>|z|)
(Intercept) 3.1704 0.9034 3.509 0.000449 ***
dist_shore -3.1710 0.7691 -4.123 3.74e-05 ***
```

```
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
Null deviance: 160.421 on 15038 degrees of freedom
Residual deviance: 48.662 on 15037 degrees of freedom
AIC: 4.3857
Number of Fisher Scoring iterations: 13
```

### A.4.2.3 GLM Output using Likelihood Ratio Tests for Selection

```
LRT Selected Model:
```

```
Call:
```

```
glm(formula = formula.glm, family = "binomial", data = complete.data.to.analyse,
weights = weights)
```

Deviance Residuals:

Min 1Q Median 3Q Max -0.5551 0.0000 0.0000 0.0000 1.3428

Coefficients:

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) 1.68802 1.67480 1.008 0.31351

dist_col -0.16160 0.11039 -1.464 0.14321

dist_shore -3.51590 0.97701 -3.599 0.00032 ***

summ_front 0.28026 0.11436 2.451 0.01426 *

spring_front -0.46366 0.19019 -2.438 0.01477 *

ss_wave 0.04123 0.01939 2.126 0.03353 *

----

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

(Dispersion parameter for binomial family taken to be 1)

Null deviance: 160.42 on 15038 degrees of freedom Residual deviance: 30.82 on 15033 degrees of freedom AIC: 12.472 Number of Fisher Scoring iterations: 14

Analysis of deviance output for reliable assessment of significance:

### A.4.2.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH_FORAGE ~ s(dist_shore, k = 3) + s(summ_front, k = 3) +
   s(spring_front, k = 3) + s(ss_wave, k = 3)
Parametric coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) -68.53 19.45 -3.524 0.000425 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
              edf Ref.df Chi.sq p-value
s(dist_shore) 1 1 11.656 0.00064 ***
s(summ_front) 1 1 4.600 0.03198 *
                     1 4.771 0.02895 *
s(spring front) 1
s(ss wave) 1
                     1 3.722 0.05369 .
```

Signif. codes: 0 `\*\*\*' 0.001 `\*\*' 0.01 `\*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.857 Deviance explained = 78.9%
REML score = 11.781 Scale est. = 1 n = 15039



Figure A.4.2.4: GAM covariate plots for Sandwich Terns in Farnes

### A.4.2.5 INLA Output

Running INLA.

Call:

```
c("inla(formula = formula.inla, family = \"binomial\", data =
complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")
```

---

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld
(Intercept)	1.3808	1.6987	-1.6641	1.2727	5.0122	0.0100
dist_shore	-3.7858	0.9796	-5.9366	-3.6982	-2.0961	0.0986
summ_front	0.2412	0.1001	0.0609	0.2352	0.4542	0.0355
spring_front	-0.3997	0.1590	-0.7426	-0.3883	-0.1185	0.0542
ss_wave	0.0359	0.0177	0.0027	0.0354	0.0723	0.0045

Model hyperparameters:

		mean	sd	0.025quant	0.5quant	0.975quant
т.0	for mesh.points-basisT	7.994	3.099	2.017	7.952	14.167

Expected number of effective parameters(std dev): 4.994(0.001232) Number of equivalent replicates : 3011.70 Marginal Likelihood: -23.58

# A.4.3 Forvie

### A.4.3.1 GLM Output using AIC for Selection

```
AIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_shore + sst_june + strat_temp,
    family = "binomial", data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q Max
-1.1334 0.0000 0.0000 0.0000 0.7863
Coefficients:
        Estimate Std. Error z value Pr(>|z|)
(Intercept) -94.133 69.710 -1.350 0.17690
dist_shore -3.254 1.138 -2.860 0.00424 **
sst_june 9.518 6.536 1.456 0.14535
strat_temp -7.090 3.533 -2.007 0.04474 *
---
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

```
(Dispersion parameter for binomial family taken to be 1)
Null deviance: 97.836 on 26516 degrees of freedom
Residual deviance: 25.004 on 26513 degrees of freedom
AIC: 9.2846
Number of Fisher Scoring iterations: 14
```

### A.4.3.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH FORAGE ~ dist shore + strat temp, family = "binomial",
   data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
   Min 1Q Median 3Q Max
-1.5768 0.0000 0.0000 0.0000 0.8352
Coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) 7.8379 3.0822 2.543 0.01099 *
dist_shore -2.7111 0.9429 -2.875 0.00404 **
strat_temp -4.6219 2.5596 -1.806 0.07096.
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
   Null deviance: 97.836 on 26516 degrees of freedom
Residual deviance: 28.799 on 26514 degrees of freedom
AIC: 8.4862
Number of Fisher Scoring iterations: 14
```

#### A.1.3.3 GLM Output using Likelihood Ratio Tests for Selection

```
LRT Selected Model:
Call:
glm(formula = formula.glm, family = "binomial", data = complete.data.to.analyse,
```

```
weights = weights)
```

```
Deviance Residuals:
```

Min	10	Median	ЗQ	Max
-1.1334	0.0000	0.0000	0.0000	0.7863

Coefficients:

```
Estimate Std. Error z value Pr(>|z|)

(Intercept) -94.133 69.710 -1.350 0.17690

dist_shore -3.254 1.138 -2.860 0.00424 **

sst_june 9.518 6.536 1.456 0.14535

strat_temp -7.090 3.533 -2.007 0.04474 *

----

Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

(Dispersion parameter for binomial family taken to be 1)

Null deviance: 97.836 on 26516 degrees of freedom Residual deviance: 25.004 on 26513 degrees of freedom AIC: 9.2846 Number of Fisher Scoring iterations: 14

#### Analysis of deviance output for reliable assessment of significance:

```
Model:
SEARCH_FORAGE ~ dist_shore + sst_june + strat_temp
Df Deviance AIC LRT Pr(>Chi)
<none> 25.004 9.285
dist_shore 1 50.042 32.323 25.0380 5.621e-07 ***
sst_june 1 28.799 11.080 3.7953 0.0513980 .
strat_temp 1 38.888 21.169 13.8845 0.0001944 ***
---
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
```

### A.1.3.4 GAM Output using Likelihood Ratio Tests for Selection

GAM Model selected (REML output):

```
Family: binomial
Link function: logit
Formula:
SEARCH FORAGE ~ s(dist shore, k = 3) + s(strat temp, k = 3)
Parametric coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) -58.15 18.89 -3.078 0.00208 **
____
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
            edf Ref.df Chi.sq p-value
s(dist_shore) 1 1 8.267 0.00404 **
s(strat_temp) 1 1 3.256 0.07118 .
____
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.324 Deviance explained = 70.6%
REML score = 8.7164 Scale est. = 1 n = 26517
```


Figure A.4.3.4: GAM covariate plots for Sandwich Terns in Forvie

## A.4.3.5 INLA Output

```
Running INLA.
Call:
c("inla(formula = formula.inla, family = \"binomial\", data =
complete.data.to.analyse, ", " weights = weights, verbose
                                        weights = weights, verbose = TRUE)")
Fixed effects:
                            sd 0.025quant 0.5quant 0.975quant
                 mean
                                                                       kld
(Intercept) 8.5092 3.0521
                                    2.9740
                                              8.3420
                                                          14.9486 0.0285
dist_shore -2.8564 0.9388
                                  -4.8381 -2.8066
                                                          -1.1474 0.0138
             -4.7950 2.5331
                                  -10.3185 -4.5855
                                                          -0.3814 0.0037
strat_temp
```

```
Model hyperparameters:
```

```
mean sd 0.025quant 0.5quant 0.975quant
T.0 for mesh.points-basisT 10.175 1.743 7.584 9.931 13.822
Expected number of effective parameters(std dev): 2.992(0.002468)
Number of equivalent replicates : 8861.99
Marginal Likelihood: -13.50
```

# A.4.4 Larne Lough

### A.4.4.1 GLM Output using AIC for Selection

```
AIC Selected Model:
Call:
glm(formula = SEARCH FORAGE ~ dist col + dist shore, family = "binomial",
   data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q Max
-1.32521 -0.00005 0.00000 0.00000 0.73625
Coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) 1.64766 0.51868 3.177 0.00149 **
dist col -0.13284 0.06301 -2.108 0.03499 *
dist_shore -3.54580 1.15139 -3.080 0.00207 **
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
   Null deviance: 210.722 on 29785 degrees of freedom
Residual deviance: 69.594 on 29783 degrees of freedom
AIC: 7.7562
```

Number of Fisher Scoring iterations: 12

### A.4.4.2 GLM Output using BIC for Selection

```
BIC Selected Model:
Call:
glm(formula = SEARCH_FORAGE ~ dist_col + dist_shore, family = "binomial",
```

```
data = complete.data.to.analyse, weights = weights)
Deviance Residuals:
    Min 1Q Median 3Q Max
-1.32521 -0.00005 0.00000 0.00000 0.73625
Coefficients:
          Estimate Std. Error z value Pr(>|z|)
(Intercept) 1.64766 0.51868 3.177 0.00149 **
dist col -0.13284 0.06301 -2.108 0.03499 *
dist shore -3.54580 1.15139 -3.080 0.00207 **
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
   Null deviance: 210.722 on 29785 degrees of freedom
Residual deviance: 69.594 on 29783 degrees of freedom
AIC: 7.7562
Number of Fisher Scoring iterations: 12
```

### A.4.4.3 GLM Output using Likelihood Ratio Tests for Selection

```
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
(Dispersion parameter for binomial family taken to be 1)
Null deviance: 210.722 on 29785 degrees of freedom
Residual deviance: 69.594 on 29783 degrees of freedom
AIC: 7.7562
Number of Fisher Scoring iterations: 12
```

#### Analysis of deviance output for reliable assessment of significance:

#### Now checking for interactions with Year:

```
Model:
SEARCH_FORAGE ~ Year * (dist_col + dist_shore)
Df Deviance AIC LRT Pr(>Chi)
<none> 65.637 19.885
Year:dist_col 2 66.433 16.681 0.79613 0.6716
Year:dist_shore 2 68.425 18.673 2.78832 0.2480
No significant Year interactions.
```

# A.4.4.4 GAM Output using Likelihood Ratio Tests for Selection

```
GAM Model selected (REML output):
Family: binomial
Link function: logit
Formula:
SEARCH FORAGE ~ s(dist col, k = 3) + s(dist shore, k = 3)
```

```
Parametric coefficients:
           Estimate Std. Error z value Pr(>|z|)
(Intercept) -29.854 8.843 -3.376 0.000735 ***
___
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
Approximate significance of smooth terms:
             edf Ref.df Chi.sq p-value
                      1 4.444 0.03503 *
s(dist col)
              1
s(dist_shore)
               1
                      1 9.481 0.00208 **
____
Signif. codes: 0 `***' 0.001 `**' 0.01 `*' 0.05 `.' 0.1 ` ' 1
R-sq.(adj) = 0.496 Deviance explained = 67%
REML score = 31.248 Scale est. = 1 n = 29786
```



Figure A.4.4: GAM covariate plots for Sandwich Terns in Larne Lough

# A.4.4.5 INLA Output

Running INLA.

#### Call:

c("inla(formula = formula.inla, family = \"binomial\", data =
complete.data.to.analyse, ", " weights = weights, verbose = TRUE)")

Fixed effects:

	mean	sd	0.025quant	0.5quant	0.975quant	kld	
(Intercept)	1.8236	0.5229	0.8811	1.7913	2.9399	0.0583	
dist_col	-0.1396	0.0630	-0.2808	-0.1325	-0.0350	0.0058	
dist_shore	-3.6656	1.1502	-6.0777	-3.5990	-1.5923	0.0059	

Model hyperparameters:

mean sd 0.025quant 0.5quant 0.975quant T.0 for mesh.points-basisT 8.157 3.191 2.050 8.091 14.568 Expected number of effective parameters(std dev): 2.998(0.00142) Number of equivalent replicates : 9934.32 Marginal Likelihood: -37.23